

Influence of natural regeneration on fractal features of residue microaggregates in bauxite residue disposal areas

by Zhu, F., Cheng, Q., Xue, S., Li, C, Hartley, W. Wu, C and Tian, T.

Copyright, Publisher and Additional Information: This is the author accepted manuscript. The final published version (version of record) is available online via Wiley. Please refer to any applicable terms of use of the publisher.

DOI: <https://doi.org/10.1002/ldr.2848>



Zhu, F., Cheng, Q., Xue, S., Li, C, Hartley, W. Wu, C and Tian, T. 2017. Influence of natural regeneration on fractal features of residue microaggregates in bauxite residue disposal areas. *Land Degradation and Development*.

17 November 2017

INTRODUCTION

Residues from mineral ore processing are disposed on land in large residue disposal areas which may eventually create a series of ecological and environmental issues (Smart *et al.*, 2016; Wu *et al.*, 2016). In the aluminum industry, bauxite residue is an alkaline solid by-product generated when alumina is extracted from bauxite ore by the Bayer process (Goloran *et al.*, 2016; Kong *et al.*, 2017a). The global inventory has reached 3.4 billion tons, with an annual increase of 120 million tons (Xue *et al.*, 2016a; Kong *et al.*, 2017b). Large volumes of bauxite residue are deposited in bauxite residue disposal areas which cause potential environmental risks, as these bare areas are sensitive to erosion by wind and water, and can be regarded as a potential source of contamination due to their high alkalinity and salinity (Gelencsér *et al.* 2011; Ruyters *et al.* 2011). In situ rehabilitation and revegetation may however stabilize the residue surface and minimize wind erosion (Courtney *et al.*, 2009; Kaur *et al.*, 2016; Schmalenberger *et al.*, 2013). Its poor physical structure is nevertheless a major limitation to support plant growth (Liu *et al.*, 2013; Zhu *et al.*, 2016a). Residue particle sizes range from 2 to 2000 μm and 60-80 % exist as $<20 \mu\text{m}$ (Xue *et al.*, 2016b). Jones *et al.* (2011) reported that addition of organic waste may influence aggregate size distribution and increase the proportion of water-stable aggregates. Zhu *et al.* (2016b) found that natural regeneration may improve the physical condition and aggregate stability of bauxite residue.

Soil aggregate stability is one of the most important properties in soils and affects water erosion, soil aeration, nutrient recycling and biological activity, as well as plant growth (Cerdà 2000; Le Guillou *et al.*, 2012; Moncada *et al.*, 2013). Physical forces, chemical bonds and biological agents may drive the aggregation processes of soil particles (Yagüe *et al.*, 2016; Lehmann & Rillig, 2015). Microaggregate stability is an important soil property which is usually used to determine soil erosion resistance (Wang *et al.*, 2016). A better understanding of microaggregate formation is essential to maintain structural stability in soils. Several major binding agents such as clay minerals, organic carbon and polyvalent ions have significant effects on colloid flocculation (Zhou *et al.*, 2005). Barbosa *et al.* (2015) observed a cementation effect by organic carbon from poultry manure applications and clay flocculation enhancing aggregation. Igwe *et al.* (2009) discovered that oxalate and pyrophosphate extractable iron-aluminum oxides may act as aggregation agents to colloidal stability; organic carbon had acted in association with the oxides as a linkage with clay particles and polyvalent cations to enhance aggregate stability. Virto *et al.* (2008) concluded that stable microaggregates were formed within the silt-size fraction and organic carbon was stored by adsorption and entrapment of fine organic residues.

Soil structure is related to the size, shape and stability of soil aggregates (Aksakal *et al.*, 2016; Ahmadi *et al.*, 2011). Microaggregate stability depends on the size distribution of microaggregates and several procedures have been proposed for characterizing aggregate size distribution. Fractal theory is mainly used to analyze the relationship between local and overall irregular broken complex images and structural geometry under different scales. The concept of fractal dimension was proposed to provide a quantitative description for irregular shapes (Jing *et al.*, 2016). Fractal theory has been widely applied in soil science to quantify and estimate aggregate size distribution of soils (Kolay & Kayabali, 2006). Fractal dimension reveals the difference between particle size distribution and related physical properties (Wei *et al.*, 2016; Wang *et al.*, 2015). Gao *et al.* (2014) suggested that fractal dimension could be regarded as a considerable and reliable parameter to reflect variations in soil properties. High

1 values represent aggregates dominated by fine fragments, whilst low values represent large fragments.
2 Many researchers have used fractal dimension to predict soil particle size distribution or the size
3 distribution of water-stable aggregates (Peng *et al.*, 2014).

4 With the development of soil fractal theory, the limitation of single-fractal dimension has been
5 stressed to describe soil particle size distribution. In order to obtain more detailed information of soil
6 structure, multi-fractal theory was introduced to soil science (Li *et al.*, 2016). Rodríguez-Lado & Lado
7 (2016) found that particle size distribution behaved as multi-fractals, with scaling properties varying in
8 different soil samples, whilst values of fractal dimension may be related to the degree of evolution of
9 the soils. Peng *et al.* (2014) found that the single- and multi-fractal parameters could describe soil
10 particle size distribution and the influences of soil structure effectively. There are, however, few studies
11 focusing on multi-fractal parameters of microaggregate size distribution.

12 This work focuses on an alumina refinery in Central China. The inventory of bauxite residue is an
13 estimated 35 million tons, which is currently increasing by approximately 2.2 million tons per annum
14 (Zhu *et al.*, 2016b). Bauxite ore is discharged in hot NaOH by the Bayer processes and the residues are
15 pumped to the disposal areas using the dry stacking method. Spontaneous vegetation colonization over
16 the past 20 years at the study site may reveal that natural weathering processes ameliorate the residue
17 substrate and support plant growth. Natural regeneration also enhances the proportion of water-stable
18 aggregates and resistance to erosion (Zhu *et al.*, 2016c). As microaggregate stability is used to predict
19 soil surface erosion (Wang *et al.*, 2016), this study focus on 1) the effect of natural weathering
20 processes on microaggregate stability of bauxite residue; 2) to evaluate microaggregate size
21 distribution in bauxite residue by fractal parameters; 3) to investigate whether fractal parameters may
22 be used as an indicator to evaluate microaggregate stability of bauxite residue.

23 MATERIALS AND METHODS

24 *Soil Sampling*

25 Residue samples were collected from a disposal area in Central China. The climate is temperate
26 continental monsoon, with a mean annual daily temperature of 12.8°C-14.8°C and average precipitation
27 ranging from 600 mm to 1200 mm per year.

28 According to ecological field investigations, five different zones related to disposal age were
29 selected during August to September 2014. These included (a) 1-year-old bauxite residue (R1), (b) 4-
30 year-old bauxite residue (R2), (c) 6-year-old bauxite residue (R3), (d) 10-year-old bauxite residue (R4),
31 and (e) 20-year-old bauxite residue (R5). Each zone was approximately 1500 m². Within the zones,
32 natural colonization only occurred in R5. For each zone, five random points, taken within 100 m x 100
33 m, were designated as the replicates. For each sampling point, the residues were sampled with an auger
34 to a depth of 20 cm. The samples were then stored in polyethylene bags, returned to the laboratory, air
35 dried at room temperature for two weeks and then subsequently passed through a 2 mm sieve prior to
36 analysis.

37 *Physical and chemical analysis*

38 Mechanical composition of residue samples were analyzed using a Malvern Mastersizer 2000
39 (Malvern Instruments Ltd., UK) (Santini and Fey, 2013). pH and electrical conductivity (EC) of residue
40 samples were determined in 1:5 solid/solution extracts. Exchangeable Ca²⁺, Mg²⁺, K⁺, and Na⁺ were
41 extracted with 1 M ammonium acetate and analyzed by ICP-AES (Jones *et al.*, 2011). Exchangeable
42

sodium percentage (ESP) was calculated as the percentage of exchangeable Na⁺ in the total exchangeable cations. The total contents of Ca, Mg, K and Na in bauxite residue were determined after microwave digestion using HF, HCl and HNO₃ and analysed by ICP-AES (Jones *et al.*, 2011). Total organic carbon was measured by the low-temperature external-heat potassium dichromate oxidation colorimetric method (Zhu *et al.*, 2016a). Chemical phases of residue samples were determined by X-ray powder diffraction (XRD) on a Bruker D8 discover 2500. XRD patterns were collected from 10° to 80° at a step size of 0.04° 2θ with a scan rate of 1° 2θ per minute and analysed using PANalytical analysis package (Zhu *et al.*, 2017).

Microaggregate Stability Analysis

Laser sizing (for the <0.25 mm fraction) was used to determine particle size distribution of residue microaggregates (Santini & Fey, 2013). In this method, 10 g of air-dried residue samples were placed in a 0.25 mm sieve. The residue samples were then immersed in distilled water and oscillated for 24 h using an end-over-end shaker with a rate of 200 cycles per minute. Particle size distribution of the <0.25 mm aggregates was determined using a Malvern Mastersizer 2000. In order to observe residue microaggregate distribution characterization under natural regeneration, micro-morphological studies of the residue microaggregates from R1 and R5 were examined using a FET Quanta-200 scanning electron microscope (SEM), equipped with energy dispersive X-ray spectroscopy. The specimen was sputter coated with a layer of gold prior to examination (Zhu *et al.*, 2016b).

Water-dispersible clay (WDC) and water-dispersible silt (WDSI) were determined as the proportion of clay and silt in suspension in the distilled water. Clay dispersion ratio (CDR) and aggregated silt+clay indices (ASC) were selected as the two indicators to measure microaggregate stability of bauxite residue. Clay dispersion ratio (CDR) was determined as the following equation (Cammeraat & Imeson, 1998):

$$CDR(\%) = \frac{\%clay + \%silt(\text{water dispersed})}{\%clay + \%silt(\text{calgon dispersed})} \times 100 \quad (1)$$

This is defined as the percentage ratio of clay+silt (<0.02 mm) obtained from both distilled water and sodium hexametaphosphate (calgon) dispersed residue samples. The value of ASC was negatively correlated to aggregate stability (Mbagwu & Auerswald, 1999).

Aggregated silt and clay (ASC) was calculated using the following equation:

$$ASC(\%) = (\%clay + \%silt)(\text{calgon dispersed}) - (\%clay + \%silt)(\text{water dispersed}) \quad (2)$$

A higher ASC value indicates greater microaggregate stability (Monreal *et al.*, 1995).

Calculation of Single-fractal Dimension (D)

The power-law relationship between either number-diameter, mass-diameter or bulk density-diameter of soil aggregates are always used to determine the fractal dimension of soil aggregates. Here, according to Tyler & Wheatcraft (1989), mass-diameter of residue aggregates was selected to calculate the fractal dimension of microaggregates, designated as D, as follows:

$$D = 3 - \lg(W_i/W_o) / \lg(\bar{d}_i/\bar{d}_{max}) \quad (3)$$

where D is the mass fractal dimension; W_i is the cumulative mass of the <d_i residue aggregates; W_o is the total mass of the residue aggregates; \bar{d}_i is the mean diameter of aggregates in adjacent particles

1 and \bar{d}_{\max} is the mean diameter of the largest aggregates.

3 Calculation of Multi-fractal Parameters

4 In this study, the measurement interval of the laser particle size analyzer ($I=[0.01 \mu\text{m}, 250 \mu\text{m}]$)
 5 was considered as the residue microaggregate size volume percentages obtained from the previous
 6 results. The microaggregate size interval is divided into 74 subintervals $I_i=[\varphi_i, \varphi_{i+1}]$, $i=1, 2, \dots, 74$.
 7 Based on standard microaggregate-size division methods, $\log(\varphi_{i+1}/\varphi_i)$ is the constant following the
 8 measurement interval of $I=[0.01, 250]$. In order to build a new measurement of the multi-fractal
 9 method, $\psi_i=\log(\varphi_i/\varphi_1)$ ($i=1, 2, \dots, 74$) was created to form a new dimensionless interval of $J=[0, 4.40]$,
 10 which had 74 subintervals of equal length, $J_i=[\psi_i, \psi_{i+1}]$ ($i=1, 2, \dots, 74$). In the interval J , ε was defined
 11 as 2^k same size subintervals, $\varepsilon=4.4 \times 2^{-k}$. The value of k ranged from 1 to 6 to make sure that every
 12 subinterval contained at least one measured value (Peng *et al.*, 2014). Thus, the multi-fractal
 13 parameters including capacity dimension (D_0), information dimension (D_1), correlation dimension (D_2)
 14 and information dimension/capacity dimension (D_1/D_0) were calculated as the following equations
 15 (Ahmadi *et al.*, 2011):

$$16 \quad D(q) \approx \lim_{\varepsilon \rightarrow 0} \frac{1}{q-1} \times \frac{\log \left[\sum_{i=1}^{n(\varepsilon)} u_i(\varepsilon)^q \right]}{\log \varepsilon} \quad q \neq 1 \quad (4)$$

$$17 \quad D_1 \approx \frac{\sum_{i=1}^{n(\varepsilon)} u_i(\varepsilon) \log u_i(\varepsilon)}{\log \varepsilon} \quad q = 1 \quad (5)$$

18 The value of q varied between -10 and 10 with a step size of 1 . The multi-fractal spectrum of the
 19 residue microaggregate size distribution were determined by $D(q)$. D_0 indicated the span of the residue
 20 microaggregate size distribution and the larger D_0 value representing a wider range; D_1 indicated the
 21 irregular degree of the residue microaggregate size distribution and the higher D_1 representing a higher
 22 level of dispersion in microaggregate size distribution; D_1/D_0 can measure the degree of heterogeneity
 23 of microaggregate size distribution (Peng *et al.*, 2014).

24 Data Analysis

25 All analyses were performed in quintuplicate. The data were statistically treated with Microsoft
 26 Excel 2003, SPSS version 19.0 and Origin 8.0. A two-way ANOVA, followed by Tukey's post hoc test
 27 was used to determine the interaction between fragment size range and residue disposal ages. Chemical
 28 properties of bauxite residue samples with different chronosequences were individually determined
 29 using one-way ANOVA followed by Tukey's post hoc tests. In the case of no homogeneity, Dunnett's
 30 T3 test was performed. Bivariate correlation analyses were used to determine the relationships between
 31 fractal parameters and residue microaggregate size distribution. All figures were constructed using
 32 Origin 8.0.

34 RESULTS AND DISCUSSION

35 Particle Size Distribution of Residue Microaggregates

36 The particle size distribution of microaggregates is shown in Table I. The main fraction, <0.02
 37 mm aggregates, accounted for more than 55% of the total microaggregate weight in newly stacked

1 residue (R1). Microaggregate fractions in the fresh residue (R1) decreased in the following order,
2 250-50 μm > 10-5 μm > 20-10 μm > 5-2 μm > 50-20 μm > <2 μm . With increasing disposal age, the
3 aggregate fraction 250-50 μm , increased significantly from 27.4% to 40.3%, whilst the clay-size
4 aggregate fraction decreased gradually. Microaggregate fractions which had been disposed for 20 years
5 (R5) decreased in the following order, 250-50 μm > 50-20 μm > 20-10 μm > 10-5 μm > 5-2 μm > <2
6 μm . With increasing disposal age, the fine particles of bauxite residue became coarser and the silt+clay
7 size (<0.02 mm) aggregate fraction effectively decreased.

8 SEM images of residue microaggregates from R1 and R5 are shown in Fig. 1. The residue
9 microaggregate in R1 contained numerous amorphous substances and fine particles. Compared to R1,
10 the microaggregate in R5 consisted of a great number of larger aggregates. With increasing disposal
11 age, the residue microaggregates had a denser structure and the particles were distributed uniformly.
12 Total sodium content decreased from 9.27% to 1.07%, whilst total calcium content increased from
13 13.89% to 27.88% (Fig. 1), which suggested that natural weathering processes decreased sodium but
14 increased calcium content. The result from Fig. 1 was consistent with the variation trend of calcium and
15 sodium content in Table II.

16 Natural weathering processes may have a positive effect on particle aggregation; the finer particles
17 aggregating to form larger particles (Santini & Fey, 2013). Climate affects soil aggregation through
18 alterations in temperature and moisture regimes and wet-dry and freeze-thaw cycles, which can
19 re-orientate particles and improve aggregation (Singer *et al.*, 1992). Weathering alters materials, which
20 are translocated within the soil through leaching, eluviation, and illuviation resulting in horizonation
21 (Garcia-Franco *et al.*, 2015; Zhou *et al.*, 2017). Plant roots and their rhizospheres have positive effects
22 on soil aggregation. Roots realign soil particles and release exudates, which result in physical, chemical
23 and biological alterations that influence aggregation (Rillig *et al.*, 2001). Courtney *et al.* (2009)
24 revealed that gypsum and spent mushroom compost application may decrease microaggregate
25 breakdown and hence the dominance of less erodible aggregates. Courtney *et al.* (2013) found that
26 addition of gypsum and compost produced a significant decrease in clay- and silt-size particles, whilst
27 mean particle size (<53 μm) was lowest in the unamended residues. Natural weathering processes
28 may ameliorate physical and chemical properties of bauxite residue, which lead to clay-size particles
29 flocculating and the formation of more stable aggregates.

31 *Microaggregate Stability of the Residues*

32 Colloidal stability indices of the residues are presented in Fig. 2. Water-dispersible clay (WDC)
33 which was used to estimate microaggregate instability, ranged from 0.64% to 2.04%, whilst
34 water-dispersible silt (WDSI), also used to estimate instability, ranged from 30.72% to 58.39%. A
35 combination of WDC and WDSI gave values of between 31.4% and 58.4%. Clay dispersion ratio
36 (CDR) of the residues ranged from 7.7% to 22.5%, and aggregated silt and clay (ASC) ranged from
37 15.3% to 19%. Following natural disposal processes, WDC, WDSI, CDR and ASC of the residues
38 stacked for 20 years (R5) increased by 218.75%, 90.07%, 192.21% and 24.18%, respectively.

39 The WDC, WDSI, CDR and ASC may be used to estimate the rate of soil dispersibility. A high
40 WDC and dispersion indices have negative implications for the soil environment in terms of water and
41 wind erosion (Virto *et al.*, 2008). With increasing disposal age, WDC, WDSI and CDR significantly
42 decreased, whilst ASC increased indicating that natural stacking processes may improve
43 microaggregate stability; the finer particles may combine together to form larger particles due to
44 binding agents related to physical and chemical properties of the residues (Zhu *et al.*, 2016c). Plant

1 growth and root penetration may also have positive effects on particle aggregation and stability as the
2 residues stacked for 20 years (R5) had improved colloidal stability compared to the other locations.

3 Stability of microaggregates, as opposed to its dispersion, is a very important soil property that
4 regulates soil degradation. Several major binding agents including clay content, organic carbon and
5 electrolytes had significant effects on microaggregate stability. Clay content, organic carbon, pH and
6 exchangeable cations were selected to identify correlational relationships with colloidal stability
7 indices. The selected physical and chemical properties were determined in a previous study (Zhu *et al.*,
8 2016c). Silt and clay contents ranged from 48.8%-23.8% and 5.9%-1.5% respectively. With increasing
9 disposal age, pH and EC were significantly reduced. Total organic carbon (TOC) content ranged from
10 5.7-10.8 g/kg. Exchangeable Ca and Na varied regularly but with opposite trends (Table II).

11 As natural weathering processes had a significant effect on bauxite residue mineral chemistry, the
12 residue samples, including R1 and R5, were selected to investigate the variation in chemical phases
13 (Figure 3). Slaked lime addition by the Bayer process resulted in the formation of calcium minerals
14 including calcite (CaCO_3), hydrogarnet ($\text{Ca}_3\text{Al}_2(\text{SiO}_4)_x(\text{OH})_{12-4x}$), and tri-calcium aluminate
15 ($\text{Ca}_3\text{Al}_2(\text{OH})_{12}$). Other alkaline minerals in the residue included sodalite ($\text{Na}_6\text{Al}_6\text{Si}_6\text{O}_{24} \cdot [2\text{NaOH}$ or
16 $\text{Na}_2\text{CO}_3]$) and cancrinite ($\text{Na}_6\text{Al}_6\text{Si}_6\text{O}_{24} \cdot 2\text{CaCO}_3$). Following natural processes, sodalite, hydrogarnet
17 and calcite decreased, as did pH. Reduction of pH resulted in the precipitation of $\text{Ca}(\text{OH})_2$ but also
18 leaching of NaOH by exchange reactions as the charged colloid (such as $\text{Al}(\text{OH})_6^{3-}$) was a regulator for
19 cation exchange (Kong *et al.*, 2017). As a result, exchangeable sodium percentage decreased. In
20 addition, sodium ions could not be coordinated with negatively charged surfaces which led to the
21 formation of alkaline dust, reduction on erosion resistance and a poor physical structure (Zhu *et al.*,
22 2016d).

23 Linear regression analysis showed that the value of clay dispersion ratio (CDR) was positively
24 correlated to clay content, pH and exchangeable Na^+ content ($r=0.898, 0.943, \text{ and } 0.826$ respectively;
25 $P<0.01$), but negatively correlated to exchangeable Ca^{2+} and total organic carbon content ($r = -0.972$
26 and $-0.936, P<0.01$) (Fig. 4). The value of aggregated silt and clay (ASC) was negatively correlated to
27 clay content, pH and exchangeable Na^+ content ($r = -0.903, -0.927, \text{ and } -0.865, \text{ respectively; } P<0.01$),
28 and positively correlated to exchangeable Ca^{2+} and total organic carbon content ($r = 0.948 \text{ and } 0.932$
29 respectively, $P<0.01$) (Fig. 5). This indicated that high exchangeable Ca^{2+} content and low
30 exchangeable Na^+ stimulated microaggregate flocculation, whilst the decrease in pH and the
31 accumulation of organic carbon may have improved microaggregate stability. Courtney *et al.* (2009)
32 established a field scale investigation to promote vegetation cover on bauxite residue, and found that
33 spent mushroom compost and gypsum amendments decreased pH and ESP which positively impacted
34 on microaggregate stability. Addition of Ca had a positive effect on flocculating clay particles, reducing
35 mechanical dispersion and lowering exchangeable Na^+ content, thereby stabilizing microaggregates
36 (Harris & Rengasamy, 2004). Pojasok & Kay (1990) reported that increasing organic carbon content
37 stimulated particle aggregation.

38 39 *Single-fractal Features of Residue Microaggregates*

40 The fractal dimension of soil microaggregates may reflect the geometry parameters of soil
41 aggregate structure, with higher clay content indicating the higher value of the fractal dimension
42 (D). Under natural weathering processes, residue fractal dimension (D) was significantly affected
43 (Fig. 6). The single-fractal dimension ranged from 2.2 to 2.4 and with increasing disposal age,
44 microaggregate fractal dimension (D) decreased. R1 had a low proportion of 250-20 μm size

1 aggregates and a high proportion of 10-2 μm aggregates, which resulted in a high fractal
2 dimension value. Under natural soil forming processes, fine particle aggregation led to the
3 decrease in single-fractal dimensions. Certainly, single-fractal dimension (D) was positively
4 correlated with the proportion of 10-5 μm , 5-2 μm , and <2 μm sized microaggregates ($r=0.859$,
5 0.977 , 0.991 respectively, $P<0.01$), but negatively correlated with the proportion of 250-50 μm
6 and 50-20 μm sized microaggregates ($r=-0.876$ and 0.761 respectively, $P<0.01$). There was no
7 significant correlation between single-fractal dimension and the proportion of 20-10 μm sized
8 microaggregates. Single-fractal dimension may be regarded as an important indicator to reflect
9 aggregate structure of bauxite residue.

10 High alkalinity and salinity resulted in poor aggregate structure of the residue and clearly
11 affected revegetation on disposal areas (Jones *et al.*, 2011). Zhu *et al.* (2016d) found that natural
12 vegetation encroachment ameliorated residue physicochemical properties and stimulated
13 aggregate stability. Courtney *et al.* (2013) investigated the physical condition of revegetated
14 residue and found that gypsum and organic carbon decreased pH and ESP, which enhanced the
15 proportion of water-stable aggregates, which supported plant growth. The related relationships
16 between single-fractal dimension and pH, EC, and ESP are displayed in Fig. 7. The single-fractal
17 dimension was positively correlated with pH, ESP, exchangeable Na^+ content and EC ($r=0.935$,
18 0.984 , 0.859 and 0.912 respectively, $P<0.01$), but negatively correlated with exchangeable Ca^{2+}
19 content ($r=-0.968$, $P<0.01$). It indicated that the single-fractal dimension of residue
20 microaggregates may reflect the related physical and chemical properties of bauxite residue.

21 22 *Multi-fractal Dimension of Residue Microaggregates*

23 Multi-fractal spectrums of residue microaggregate size distributions between -10 and 10 at 1.0 lag
24 increments for different disposal ages are presented in Fig. 8. The multi-fractal spectrums show a
25 typical anti-S-decreasing function. The information entropy ($D_{(q)}$) of residue microaggregates
26 decreased with increasing disposal age. Furthermore, in each residue sample, $D_0>D_1$ always existed,
27 meaning that microaggregate size distribution with different disposal ages were not homogeneous or
28 monofractal. Therefore, multi-fractal dimension analysis was essential.

29 The value of D_0 , D_1 and D_1/D_0 decreased with increasing disposal age (Table III). In residues
30 which had been stacked for 20 years (R5), these values were nearly the lowest (D_0 , D_1 and D_1/D_0
31 values of 0.942 , 0.853 and 0.906 , respectively), whilst for R1, these values were the highest (D_0 , D_1
32 and D_1/D_0 values of 0.968 , 0.891 and 0.920 , respectively). Analysis of variance in the different residues
33 showed that the multi-fractal parameters of R1 and R2 were significantly different ($P<0.05$).

34 The larger D_0 means a wider range of microaggregate size distributions. Nevertheless, the
35 calculation of D_0 is based on the assumption that particle size distribution was homogeneous. The
36 value of D_1/D_0 may make a quantitative description of the heterogeneous degree of soil particle
37 size distribution. Miranda *et al.* (2006) pointed out that if the value of D_1/D_0 was closer to 1 this
38 specified that the particle size distribution was more concentrated. The value of D_1/D_0 in the
39 residue microaggregates ranged from 0.896 to 0.920 . With increasing disposal age, the value of
40 D_1/D_0 decreased indicating that natural weathering processes decreased the concentration of
41 microaggregate distribution. Natural processes accumulate organic carbon over time and this may
42 have ameliorated the high alkalinity and salinity in the residue, stimulating fine particle
43 aggregation. A significant increase in the proportion of 250-50 μm residue microaggregates resulted
44 in homogeneity with increasing disposal age.

1 Bivariate correlation analysis between multi-fractal parameters and residue microaggregate
2 distribution showed that the value of D_0 , D_1 , D_1/D_0 was significantly correlated with the
3 proportion of $<2 \mu\text{m}$ microaggregates ($r=0.915$, 0.786 and 0.523 respectively, $P<0.05$). In addition, the
4 value of D_1 was positively correlated with the proportion of $10\text{-}5 \mu\text{m}$ and $5\text{-}2 \mu\text{m}$ microaggregates
5 ($r=0.912$ and 0.671 respectively, $P<0.05$). According to correlation analysis between multi-fractal
6 parameters, D_1 was positively correlated with D_0 and D_1/D_0 ($r=0.933$ and 0.917 respectively, $P<0.05$).
7 Multi-fractal parameters of residue microaggregate size distribution were mainly affected by the
8 proportion of $<10 \mu\text{m}$ microaggregates, especially the silt-sized ($<2 \mu\text{m}$) microaggregates.

9 Multi-fractal dimension of residue microaggregates was closely linked with related
10 physicochemical properties. Residues which had been stacked for 20 years had a low pH and EC which
11 led to a low multi-fractal dimension, whilst the newly stacked residue had a high pH and EC and a high
12 multi-fractal dimension. This suggested that the multi-fractal parameters of microaggregates may
13 reflect physical and chemical properties and may be used as an effective indicator to characterize
14 alkalinity and salinity of bauxite residue. According to multiple linear models between multi-fractal
15 parameters and the related properties of bauxite residue, the following equations were obtained:

$$16 \quad D_0=1.01926+0.0018x_1-0.0079x_2-0.00376x_3+2.06\times 10^{-4}x_4 \quad (6)$$

$$17 \quad D_1=0.91468+0.00325x_1-0.00476x_2+1.9669\times 10^{-4}x_3-3.91826\times 10^{-4}x_4 \quad (7)$$

$$18 \quad D_1/D_0=0.88054+0.00498x_1-0.0019x_2+0.00231x_3+6.58484\times 10^{-5}x_4 \quad (8)$$

19 where x_1 is the content of TOC, x_2 is the value of pH, x_3 is the value of EC, and x_4 is the value of ESP.
20 This demonstrated that organic carbon content and pH were the main properties influencing the values
21 of multi-fractal dimension of residue microaggregates.

22 *Relationship between Microaggregate Stability and Fractal Parameters*

23 Microaggregate stability is usually used to estimate or predict soil erosion and surface runoff
24 (Wang *et al.*, 2016). Bauxite residue has poor physical structure to resist water erosion and support
25 revegetation. Zhu *et al.* (2016d) discovered that following natural weathering processes, the erodibility
26 factor of the residue decreased indicating improved resistance to erosion. Correlation analysis showed
27 that single- and multi-fractal parameters were significantly correlated to microaggregate stability
28 indicating that fractal parameters of residue microaggregate distribution may reflect microaggregate
29 stability. The fractal dimension of microaggregate size distribution may exhibit variation in
30 microaggregate size distribution. The high value fractal parameters indicated a dense physical structure
31 and poor erosion resistance.

32 Tang *et al.* (2013) revealed a significant negative relationship between fractal dimension and soil
33 microaggregate content ($<0.25 \text{ mm}$) in karst rocky desertification areas and suggested that fractal
34 dimension could be used as a reliable indicator of soil quality. A small fractal dimension value for
35 granular structure indicated a stable soil structure. Ahmadi *et al.* (2011) found that both number- and
36 mass-based fragmentation fractal dimension may describe the aggregate size distribution and estimate
37 splash and inter-rill soil erosion. In our study, the single-fractal dimension (D) of microaggregate
38 distribution was negatively correlated with ASC ($r=-0.977$, $P<0.01$), whilst positively correlated with
39 CDR ($r=0.995$, $P<0.01$). The value of D_0 and D_1 showed a significant correlation with ASC ($r=-0.823$
40 and -0.739 respectively, $P<0.01$) and CDR ($r=0.822$ and 0.709 respectively, $P<0.01$), whilst D_0/D_1 had
41

1 little significant difference with ASC and CDR (Table IV). With increasing disposal age, aggregation
2 of fine particles resulted in a lower value of fractal dimension and a more stable aggregated structure.
3 This suggests that fractal dimension may be useful to characterize microaggregate stability. Compared
4 to multi-fractal parameters, single-fractal dimension (D) was more significantly correlated with
5 microaggregate stability in bauxite residue.

6 CONCLUSIONS

7 Microaggregate stability, an important physical indicator, is required to sustain a stable physical
8 structure. This study has clearly demonstrated that natural weathering processes significantly affect
9 particle size distribution of residue microaggregates. With increasing disposal age, the proportion of
10 silt- and clay-sized microaggregates significantly decreased. Clay dispersion ratio (CDR) decreased
11 from 22.5% to 7.7%, and aggregated silt and clay (ASC) increased from 15.3% to 19% indicating that
12 natural weathering processes enhanced microaggregate stability. Clay content, organic carbon,
13 exchangeable bases and pH were significantly correlated with ASC and CDR which indicated that
14 organic carbon and exchangeable cations had significant effects on microaggregate stability. The value
15 for single-fractal dimension (D) varied from 2.2 to 2.4. With increasing disposal age, both single-fractal
16 dimension (D) and multi-fractal parameters (D_0 , D_1 and D_1/D_0) decreased, revealing that natural
17 weathering process promoted aggregation of microaggregates. Correlation analyses demonstrated that
18 fractal parameters were significantly correlated with microaggregate stability and physicochemical
19 properties, indicating that fractal parameters may be used to characterize residue physical structure and
20 related properties. This study may help to provide an improved understanding of physical
21 microstructures, appropriate indicators to use when evaluating microaggregate stability, and a scientific
22 basis for the revegetation of bauxite residue disposal areas.

23 ACKNOWLEDGEMENT

24 Financial support from National Natural Science Foundation of China (No. 41371475) and
25 Environmental protection's special scientific research for Chinese public welfare industry (No.
26 201509048) is gratefully acknowledged.

27 REFERENCES

- 28
- 29 Ahmadi A, Neyshabouri M, Rouhipour H, Asadi H. 2011. Fractal dimension of soil aggregates as an index of soil erodibility.
30 *Journal of Hydrology* **400**: 305-311. DOI: 10.1016/j.jhydrol.2011.01.045.
- 31 Aksakal EL, Sari S, Angin I. 2016. Effects of vermicompost application on soil aggregation and certain physical properties. *Land*
32 *Degradation & Development* **18**: 1916-1932. DOI:10.1002/ldr.2350.
- 33 Barbosa GMDC, Oliveira JFD, Miyazawa M, Ruiz DB, Filho JT. 2015. Aggregation and clay dispersion of an oxisol treated with
34 swine and poultry manures. *Soil and Tillage Research* **146**: 279-285. DOI: 10.1016/j.still.2014.09.022.
- 35 Cammeraat LH, Imeson AC. 1998. Deriving indicators of soil degradation from soil aggregation studies in southeastern Spain
36 and southern France. *Geomorphology* **23**: 307-321. DOI: 10.1016/S0169-555X(98)00012-9.
- 37 Cerdà A. 2000. Aggregate stability against water forces under different climates on agriculture land and scrubland in southern

1 Bolivia. *Soil & Tillage Research* **57**: 159-166. DOI: 10.1016/S0167-1987(00)00155-0.

2 Courtney RG, Jordan SN, Harrington T. 2009. Physico-chemical changes in bauxite residue following application of spent
3 mushroom compost and gypsum. *Land Degradation & Development* **20**: 572-581. DOI: 10.1002/ldr.926.

4 Courtney R, Harrington T, Byrne KA. 2013. Indicators of soil formation in restored bauxite residues. *Ecological Engineering* **58**:
5 63-68. DOI: 10.1016/j.ecoleng.2013.06.022.

6 García-Franco N, Martínez-Mena M, Goberna M, Albaladejo J. 2015. Changes in soil aggregation and microbial community
7 structure control carbon sequestration after afforestation of semiarid shrublands. *Soil Biology & Biochemistry* **87**: 110-121.
8 DOI: 10.1016/j.soilbio.2015.04.012.

9 Gao GL, Ding GD, Zhao YY, Wu B, Zhang YQ, Qin SG, Bao YF, Yu MH, Liu YD. 2014. Fractal approach to estimating
10 changes in soil properties following the establishment of *Caragana korshinskii*, shelterbelts in Ningxia, NW China.
11 *Ecological Indicators* **43**: 236-243. DOI: 10.1016/j.ecolind.2014.03.001.

12 Gelencsér A, Kováts N, Turóczy B, Rostási A, Hoffer A, Imre K, Nyirő-Kósa I, Csákberényi-Malasics D, Tóth A, Czitrovsky A,
13 Nagy A, Nagy S, Ács A, Kovács A, Ferincz A, Hartványi Z, Pósfai M. 2011. The red mud accident in Ajka (Hungary):
14 characterization and potential health effects of fugitive dust. *Environmental Science & Technology* **45**: 1608-1615. DOI:
15 10.1021/es104005r.

16 Goloran JB, Phillips IR, Chen CR. 2016. Forms of nitrogen alter plant phosphorus uptake and pathways in rehabilitated highly
17 alkaline bauxite processing residue sand. *Land Degradation & Development*. DOI: 10.1002/ldr.2630.

18 Rodríguez-Lado L, Lado M. 2016. Relation between soil forming factors and scaling properties of particle size distributions
19 derived from multifractal analysis in topsoils from Galicia (NW Spain). *Geoderma*. DOI: 10.1016/j.geoderma.2016.08.005.

20 Harris MA, Rengasamy P. 2004. Sodium affected subsoils, gypsum, and green-manure: Inter- actions and implications for
21 amelioration of toxic red mud wastes. *Environmental Geology* **45**: 1118-1130. DOI: 10.1007/s00254-004-0970-y.

22 Igwe CA, Zarei M, Stahr K. 2009. Colloidal stability in some tropical soils of southeastern Nigeria as affected by iron and
23 aluminium oxides. *Catena* **77**: 232-237. DOI: 10.1016/j.catena.2009.01.003.

24 Jing J, Feng P, Wei S, Zhao H, Liu Y. 2016. Investigation on the surface morphology of Si_3N_4 ceramics by a new fractal
25 dimension calculation method. *Applied Surface Science* **387**: 812-821. DOI: 10.1016/j.apsusc.2016.06.181.

26 Jones BEH, Haynes RJ, Phillips IR. 2011. Influence of organic waste and residue mud additions on chemical, physical and
27 microbial properties of bauxite residue sand. *Environmental Science and Pollution Research* **18**: 199-211. DOI:
28 10.1007/s11356-010-0364-5.

29 Kaur N, Phillips I, Fey MV. 2016. Amelioration of bauxite residue sand by intermittent additions of nitrogen fertiliser and
30 leaching fractions: the effect on growth of kikuyu grass and fate of applied nutrients. *Science of the Total Environment* **550**:
31 362-371. DOI: 10.1016/j.scitotenv.2016.01.012.

32 Kong XF, Guo Y, Xue SG, William H, Wu C, Ye YZ, Cheng QY. 2017a. Natural evolution of alkaline characteristics in bauxite
33 residue. *Journal of Cleaner Production* **143**: 224-230. DOI: 10.1016/j.jclepro.2016.12.125.

34 Kong XF, Li M, Xue SG, William H, Chen CR, Wu C, Li XF, Li YW. 2017b. Acid transformation of bauxite residue: conversion
35 of its alkaline characteristics. *Journal of Hazardous Materials* **324**: 382-390. DOI: 10.1016/j.jhazmat.2016.10.073.

36 Le Guillou C, Angers DA, Maron PA, Leterme P, Menasseri-Aubry S. 2012. Linking microbial community to soil water-stable
37 aggregation during crop residue decomposition. *Soil Biology and Biochemistry* **50**: 126-133. DOI:
38 10.1016/j.soilbio.2012.03.009.

39 Lehmann A, Rillig MC. 2015. Understanding mechanisms of soil biota involvement in soil aggregation: a way forward with
40 saprobic fungi? *Soil Biology and Biochemistry* **88**: 298-302. DOI: 10.1016/j.soilbio.2015.06.006.

41 Li B, Liu R, Jiang Y. 2016. A multiple fractal model for estimating permeability of dual-porosity media. *Journal of Hydrology*
42 **540**: 659-669. DOI: 10.1016/j.jhydrol.2016.06.059.

43 Liu Y, Naidu R, Ming H. 2013. Surface electrochemical properties of red mud (bauxite residue): zeta potential and surface
44 charge density. *Journal of Colloid and Interface Science* **394**: 451-457. DOI: 10.1016/j.jcis.2012.11.052.

- 1 Mbagwu JSC, Auerswald K. 1999. Relationship of percolation stability of soil aggregates to land use, selected properties,
2 structural indices and simulated rainfall erosion. *Soil & Tillage Research* **50**: 197-206. DOI:
3 10.1016/S0167-1987(99)00006-9.
- 4 Miranda JGV, Montero E, Alves MC, González AP, Vázquez EV. 2006. Multifractal characterization of saprolite particle-size
5 distributions after topsoil removal. *Geoderma* **134**: 373-385. DOI: 10.1016/j.geoderma.2006.03.014.
- 6 Moncada MP, Gabriels D, Cornelis W, Lobo D. 2013. Comparing aggregate stability tests for soil physical quality indicators.
7 *Land Degradation & Development* **26**: 843-852. DOI: 10.1002/ldr.2225.
- 8 Monreal CM, Schnitzer M, Schulten HR, Campbell CA, Anderson DW. 1995. Soil organic structures in macro and
9 microaggregates of a cultivated Brown Chernozem. *Soil Biology and Biochemistry* **27**: 845-853. DOI:
10 10.1016/0038-0717(94)00220-U.
- 11 Peng G, Xiang N, Lv S, Zhang G. 2014. Fractal characterization of soil particle-size distribution under different land-use patterns
12 in the Yellow River Delta Wetland in China. *Journal of Soils and Sediments* **14**: 1116-1122. DOI:
13 10.1007/s11368-014-0876-6.
- 14 Pojasok T, Kay BD. 1990. Assessment of a combination of wet sieving and turbidimetry to characterize the structural stability of
15 moist aggregates. *Canadian Journal of Soil Science* **70**: 33-42. DOI: 10.4141/cjss90-004.
- 16 Rillig MC, Wright SF, Kimball BA, Pinter PJ, Wall GW, Ottman MJ, Leavitt SW. 2001. Elevated carbon dioxide and irrigation
17 effects on water stable aggregates in a Sorghum field: a possible role for arbuscular mycorrhizal fungi. *Global Change*
18 *Biology* **7**: 333-337. DOI: 10.1046/j.1365-2486.2001.00404.x.
- 19 Ruyters S, Mertens J, Vassilieva E, Dehandschutter B, Poffijn A, Smolders E. 2011. The red mud accident in Ajka (Hungary):
20 plant toxicity and trace metal bioavailability in red mud contaminated soil. *Environmental Science & Technology* **45**:
21 1616-1622. DOI: 10.1021/es104000m.
- 22 Santini TC, Fey MV. 2013. Spontaneous vegetation encroachment upon bauxite residue (red mud) as an indicator and facilitator
23 of in situ remediation processes. *Environmental Science & Technology* **47**: 12089-12096. DOI: 10.1021/es402924g.
- 24 Schmalenberger A, O Sullivan O, Gahan J, Cotter PD, Courtney R. 2013. Bacterial communities established in bauxite residues
25 with different restoration histories. *Environmental Science & Technology* **47**: 7110-7119. DOI: 10.1021/es401124w.
- 26 Smart D, Callery S, Courtney R. 2016. The potential for waste-derived materials to form soil covers for the restoration of mine
27 tailings in Ireland. *Land Degradation & Development* **27**: 542-549. DOI: 10.1002/ldr.2465.
- 28 Kolay E, Kayabali K. 2006. Investigation of the effect of aggregate shape and surface roughness on the slake durability index
29 using the fractal dimension approach. *Engineering Geology* **86**: 271-284. DOI: 10.1016/j.enggeo.2006.05.007.
- 30 Singer MJ, Southard RJ, Warrington DN, Janitzky P. 1992. Stability of synthetic sand-clay aggregates after wetting and drying
31 cycles. *Soil Science Society of America Journal* **56**: 1843-1848. DOI:10.2136/sssaj1992.03615995005600060032x.
- 32 Tang Y, Li J, Zhang X, Yang P, Wang J, Zhou N. 2013. Fractal characteristics and stability of soil aggregates in karst rocky
33 desertification areas. *Natural Hazards* **65**: 563-579. DOI: 10.1007/s11069-012-0383-2.
- 34 Tyler SW, Wheatcraft SW. 1989. Application of fractal mathematics to soil water retention estimation. *Soil Science Society of*
35 *America Journal* **53**: 987-996. DOI:10.2136/sssaj1989.03615995005300040001x.
- 36 Virto I, Barré P, Chenu C. 2008. Microaggregation and organic matter storage at the silt-size scale. *Geoderma* **146**: 326-335.
37 DOI: 10.1016/j.geoderma.2008.05.021.
- 38 Wang J, Yang W, Yu B, Li Z, Cai C, Ma R. 2016. Estimating the influence of related soil properties on macro- and
39 micro-aggregate stability in ultisols of south-central China. *Catena* **137**: 545-553. DOI: 10.1016/j.catena.2015.11.001.
- 40 Wang J, Zhang M, Bai Z, Guo L. 2015. Multi-fractal characteristics of the particle distribution of reconstructed soils and the
41 relationship between soil properties and multi-fractal parameters in an opencast coal-mine dump in a loess area.
42 *Environmental Earth Sciences* **73**: 4749-4762. DOI: 10.1007/s12665-014-3761-0.
- 43 Wei X, Li X, Wei N. 2016. Fractal features of soil particle size distribution in layered sediments behind two check dams:
44 Implications for the Loess Plateau, China. *Geomorphology* **266**: 133-145. DOI: 10.1016/j.geomorph.2016.05.003.

1 Wu C, Zou Q, Xue SG, Pan WS, Yue X, William H, Huang L, Mo JY. 2016. Effect of silicate on arsenic fractionation in soils
2 and its accumulation in rice plants. *Chemosphere* **165**: 478-486. DOI: 10.1016/j.chemosphere.2016.09.061.

3 Xue SG, Kong XF, Zhu F, William H, Li XF, Li YW. 2016a. Proposal for management and alkalinity transformation of bauxite
4 residue in China. *Environmental Science and Pollution Research* **23**: 12822-12834. DOI: 10.1007/s11356-016-6478-7.

5 Xue SG, Zhu F, Kong XF, Wu C, Huang L, Huang N, William H. 2016b. A review of the characterization and revegetation of
6 bauxite residues (Red mud). *Environmental Science and Pollution Research* **23**: 1120-1132. DOI:
7 10.1007/s11356-015-4558-8.

8 Yagüe MR, Domingo-Olivé F, Bosch-Serra ÀD, Poch RM, Boixadera J. 2016. Dairy cattle manure effects on soil quality:
9 porosity, earthworms, aggregates and soil organic carbon fractions. *Land Degradation & Development*.
10 DOI:10.1002/ldr.2477.

11 Zhou Q, Bao Y, Liu W. 2017. *Ecological Geoscience*. Beijing: Science Press (in Chinese).

12 Zhou Q, Sun F, Liu R. 2005. Joint chemical flushing of soils contaminated with petroleum hydrocarbons. *Environment*
13 *International* **31**: 835-839. DOI: 10.1016/j.envint.2005.05.039.

14 Zhu F, Huang N, Xue SG, William H, Li YW, Zou Q. 2016a. Effects of binding materials on microaggregate size distribution in
15 bauxite residues. *Environmental Science and Pollution Research* **23**: 23867-23875. DOI: 10.1007/s11356-016-7626-9.

16 Zhu F, Liao JX, Xue SG, William H, Zou Q, Wu H. 2016b. Evaluation of aggregate microstructures following natural
17 regeneration in bauxite residue as characterized by synchrotron-based X-ray micro-computed tomography. *Science of The*
18 *Total Environment* **573**: 155-163. DOI: 10.1016/j.scitotenv.2016.08.108.

19 Zhu F, Xue SG, William H, Huang L, Wu C, Li XF. 2016c. Novel predictors of soil genesis following natural weathering
20 processes of bauxite residues. *Environmental Science and Pollution Research* **23**: 2856-2863. DOI:
21 10.1007/s11356-015-5537-9.

22 Zhu F, Zhou JY, Xue SG, William H, Wu C, Guo Y. 2016d. Aging of bauxite residue in association of regeneration: a
23 comparison of methods to determine aggregate stability & erosion resistance. *Ecological Engineering* **92**: 47-54. DOI:
24 10.1016/j.ecoleng.2016.03.025.

25 Zhu F, Hou JT, Xue SG, Wu C, Wang QL, William H. 2017. Vermicompost and gypsum amendments improve aggregate
26 formation in bauxite residue. *Land Degradation & Development*. DOI: 10.1002/ldr.2737.

27
28
29