

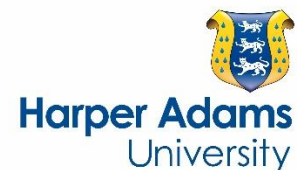
The diet of Eurasian Tree Sparrow *Passer montanus* nestlings in relation to agri-environment scheme habitats

by McHugh, N.H., Prior, M., Leather, S.R. and Holland, J.M.

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McHugh, N.H., Prior, M., Leather, S.R. and Holland, J.M. 2016. The diet of Eurasian Tree Sparrow *Passer montanus* nestlings in relation to agri-environment scheme habitats. *Bird Study*, 63 (2), pp. 279-283.

1 **The diet of Eurasian Tree Sparrow *Passer montanus***
2 **nestlings in relation to agri-environment scheme habitats**
3

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15 **Capsule** It has been suggested by some authors that the UK agri-environment ‘wild bird
16 seed’ option negatively impacts Tree Sparrow populations in the UK. Here we provide
17 evidence for a change in nestling diet with increasing wild bird seed coverage and propose a
18 possible mechanism for its negative impact on population trends.

19

20 The intensification of agriculture has been implicated as a major factor driving the population
21 decline of farmland birds including the Eurasian Tree Sparrow *Passer montanus* (hereafter
22 Tree Sparrow) in the United Kingdom (UK; Newton 2004). The Tree Sparrow is a mixed
23 diet species; adults require grain and wild plant seed but nestlings are dependent on
24 invertebrate food resources (Holland *et al.* 2006). Across Europe, farmland invertebrate
25 populations have decreased due to the increased use of pesticides and herbicides (Stoate *et al.*
26 2001). Additionally, the proportion of non-cropped areas available to foraging birds have
27 declined (Stoate *et al.* 2001). Insect taxa are an essential protein source for farmland bird
28 chicks and reduced invertebrate availability may have detrimental consequences on chick
29 survival, affecting their development and flight feather growth (Borg & Toft 1999, 2000;
30 Southwood *et al.* 2002) as well as increasing their risk of hypothermia (Potts 2012). When
31 invertebrates are scarce, farmland birds such as Yellowhammer *Emberiza citrinella* and Cirl
32 Bunting *Emberiza cirulus*, are known to supplement nestling diet with seed despite its lower
33 protein and energy content to the equivalent weight of invertebrates (Evans *et al.* 1997;
34 Douglas *et al.* 2009).

35 Agri-environment schemes (AES) comprise a suite of prescriptive management
36 strategies that are employed across Europe to, in part, alleviate biodiversity problems related
37 to agricultural intensification (Kleijn & Sutherland 2003) The English AES, Environmental
38 Stewardship (ES) ~~contained~~ offered several habitat options that should boost Tree Sparrow
39 chick food availability, including ungrazed grass margins and field corners (Vickery *et al.*

40 2002). In contrast, the value of an ES wild bird seed (WBS) option to breeding Tree Sparrow
41 is currently the subject of debate. WBS is designed as a seed-rich food resource for
42 granivorous birds in winter. Holland *et al.* (2014) showed that at a plot scale this habitat can
43 also provide high levels of chick food for farmland birds during the breeding season, however
44 this calculation included some invertebrate groups that are uncommon in the diet of Tree
45 Sparrow nestlings e.g. Nuroptera and Formicidae (Field *et al.* 2008). More recently, Bright
46 *et al.* (2015) ~~has~~ reported regional scale declines in breeding densities of Tree Sparrow
47 relative to the area of seed-rich habitat available, a finding ~~that was~~ consistent with Baker *et*
48 *al.* (2012) who described a negative relationship between Tree Sparrow population growth
49 and the area of WBS on mixed farmland. High concentrations of feeding birds leading to
50 increased predation pressure was the suggested cause of this negative effect (Baker *et al.*
51 2012), but here we investigate an alternative mechanism for declining populations by relating
52 nestling diet to the prevalence of this habitat.

53 The aim of this study was to define the dietary niche of Tree Sparrow nestlings and to
54 investigate if the presence of key invertebrate food items or seed in their diet is influenced by
55 the coverage of grass AES habitat (an aggregate group consisting of a number of structurally-
56 similar grassy habitats such as grass margins and wildflower margins) or annual WBS ES
57 habitats on arable farmland. The following predictions were tested: (1) The presence of key
58 invertebrate food groups were expected to positively correlate with Grass AES coverage and
59 (2) The presence of seed in faecal sacs were expected to positively correlate with WBS cover.

60 From mid-June to July 2013, nestling diet on 17 Tree Sparrow colony sites (from 9
61 farms) on the Marlborough and Pewsey Downs was assessed (Figure 1). This area has been
62 designated as high priority for Entry Level Stewardship farmland bird conservation by
63 Natural England. Sites were mixed farmland with habitat types available to colonies
64 including permanent pasture (18 883.461±3116.256m²), arable crops (92

65 654.1503±3028.375m²; barley, *Triticum*, wheat, *Hordeum* and oilseed rape, *Brassica napus*
66 *spp.*) along with small patches of woodland (1682.962±358.403m²). Nestling diet was
67 assessed from faecal samples (n=83) collected from 41 broods where nestlings were between
68 7 and 10 days old. This represents a period when chicks develop rapidly and energy is being
69 invested in feather growth (Ramsay & Houston 2003). Samples were stored in tubes and
70 frozen before being processed for identification. Faecal analysis was used to define Tree
71 Sparrow diet following the method described by Moreby (1988). The presence of seed and
72 cereal husks in samples was also recorded and grouped under the category “seed”.

73 We analysed how nestling diet relates to grass ES (mean ± SE= 1898.533±308.344
74 m²; range = 0-18 222 m²) and WBS (mean ± SE =1452.027±239.452 m²; range =0-5026.536
75 m²) habitat coverage within the average foraging range of an adult Tree Sparrow (200m;
76 Summer-Smith, 1995). Using Generalized Linear Mixed-effects Models (GLMMs) with the
77 packages lme4 and language R, in R version 3.0.3 (Bates *et al.* 2015; R Core Development
78 Team, 2014) the response variables were: 1. Presence or absence of taxon groups comprising
79 >5% nestling diet (see later); 2. The presence/absence of seed in faecal sacs. Faecal analysis
80 may underrepresent species identified by fragile structures that are often completely digested
81 by the animal and over-report those identified by more robust remains (Gooch *et al.* 2015).
82 Because of this, data on the percentage occurrence of key food items were not analysed as no
83 corrections factor specific to Tree Sparrow exist that account for the possible undercounting
84 of soft bodied food items.

85 Farms, colonies within farms and a brood identification number were included in
86 models as nested random effects. GLMMs were constructed with a binomial error
87 distribution and logit link function. The package LMERConvenienceFunctions was used to
88 check model assumptions (Tremblay 2015).

89 All Tree Sparrow nestling faecal samples contained invertebrate remains, comprising
90 Araneae ($7.45 \pm 0.90\%$ of all invertebrate food items), Carabidae ($16.41 \pm 1.54\%$), other
91 adult Coleoptera (Cantharidae, Chrysomelidae, Coccinellidae, Curculionidae, Elateridae,
92 Staphylinidea, Scarabidae; $15.32 \pm 1.69\%$), Coleoptera larvae ($14.19 \pm 2.41\%$), Diptera
93 ($22.06 \pm 1.60\%$), Lepidoptera Larvae ($6.29 \pm 1.46\%$), Tipulidae ($11.27 \pm 0.50\%$) and other
94 invertebrates (Acarina, Aphididae, Dermaptera, Gastropoda, Homoptera, Hymenoptera,
95 Opiliones, unidentified Coleoptera; $7.01 \pm 0.98\%$) and seed was present in 51% of faecal
96 samples (n=83) and was fed to 78% of broods (n=41). Faecal sacs were more likely to
97 contain seed where WBS coverage was high, but had no significant relationship with grass
98 ES (Table 1). No correlations between the invertebrate taxa investigated and grass ES or
99 WBS coverage were found (Table 1). It is important to consider that because this study
100 involved multiple statistical tests, it is possible that some of the observed effects are type I
101 errors.

102 Past studies of Tree Sparrow diet have highlighted Lepidoptera as a major dietary
103 component (approximately 28%; Holland *et al.* 2006). In this study, however, Lepidoptera
104 larvae accounted for only 6.29% of their diet. This finding may reflect national declines in
105 Lepidoptera abundance, a theory that has been proposed by Field *et al.* (2008), who found
106 Lepidoptera only represented 7% of Tree Sparrow chick food items. There is evidence that
107 nationally Lepidoptera have declined over the same period as threatened farmland bird
108 species (Benton *et al.* 2002; Conrad *et al.* 2006; Fox *et al.* 2011).

109 Although the invertebrate taxa consumed by Tree Sparrow chicks were unaffected by
110 grass ES coverage the presence of grain in their diet positively correlated with WBS
111 coverage. Invertebrate food provides a better source of protein and supplies particular amino
112 acids that facilitate growth; these are often absent or only present in very low proportions in
113 plant food (Potts 2012). This is known to depress nestling body condition in other farmland

114 bird species e.g. Yellowhammer (Douglas *et al.* 2012) and can impact their future survival
115 and fitness as a consequence (Wright *et al.* 1998, Lindstrom 1999).

116 WBS is primarily a winter habitat and was represented by short ($0.35\text{m} \pm 0.22\text{m}$)
117 sparse vegetation at the time of sampling (pers obs). Invertebrate abundance increases with
118 the height and structural diversity of a habitat (Eyre & Leifert 2011) and it is therefore
119 unlikely that invertebrate food resources were abundant in this habitat. WBS is generally
120 planted in April or May meaning that during the peak breeding season (May-July) the habitat
121 is not sufficiently developed to provide seeds for foraging adults. Since spring sown WBS
122 appears to provide little in the way of food during the breeding season, Tree Sparrows may be
123 resorting to feeding in cropped areas instead, and as they support few insects (Holland *et al.*
124 2012), this is responsible for the higher prevalence of grain in nestling diets. This does not
125 necessarily negate the benefits of WBS as a winter food resource (Stoate *et al.* 2004), but it is
126 important that it does not come at the cost of brood rearing resources that are vital to maintain
127 productivity. WBS may be improved as a summer foraging habitat by sowing in the autumn
128 instead of spring, this practice is already carried out by some farmers and results in a more
129 mature spring/summer crop which should ~~positively impact~~ result in increased invertebrate
130 populations. Planting two year in place of annual WBS strips may also benefit breeding Tree
131 Sparrow as two-year strips are much better at providing invertebrates in their second year due
132 to increased weed cover (J. Holland *et al.* unpubl. data).

133 The increased presence of grain in the diet of nestlings with WBS coverage may offer
134 an explanation for declining Tree Sparrow population growth on mixed farmland, but it
135 assumes this relationship reflects a decision by parents to supplement nestling diet with grain
136 at the cost of invertebrates. Further research is needed in order to verify that increased seed
137 intake results in reduced insect mass within the diet but this is currently limited as no
138 correction factors for Tree Sparrow faecal analysis were available to account for potentially

139 undercounting soft bodied prey. Correction factors may also be important in investigation the
140 relationship between the abundance of key dietary items and grass AES as the
141 presence/absence data used in our analysis may have been too coarse to detect such a
142 relationship.

143

144 **Acknowledgements**

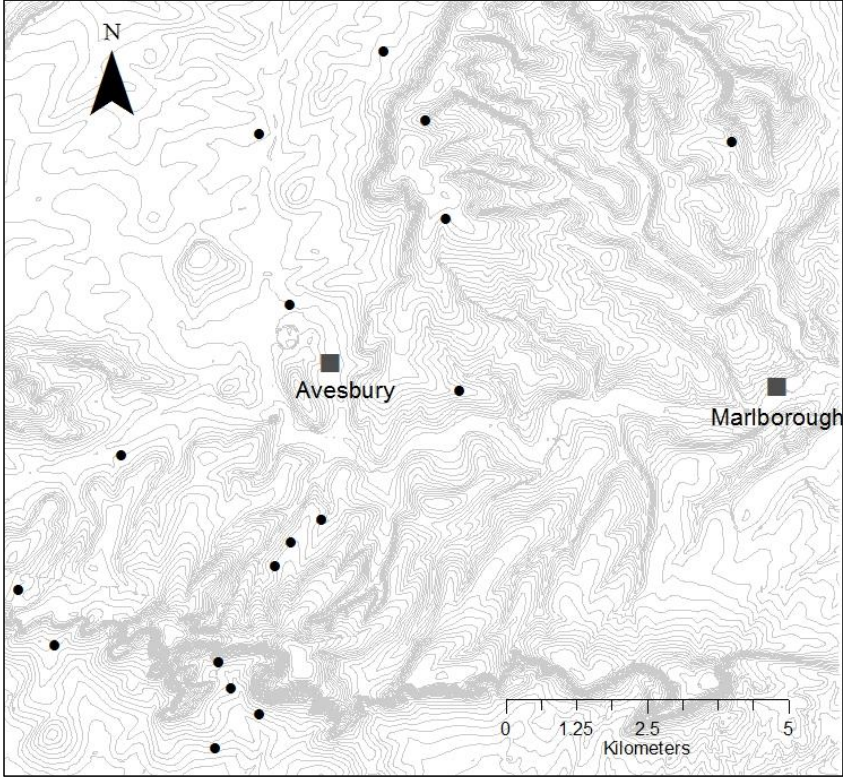
145 This project was funded by a BBSRC case studentship with the Game and Wildlife
146 Conservation Trust (GWCT), additional funding was provided by Natural England. We are
147 grateful to the landowners who granted us access to their land over the course of the project.
148 Thanks to CJ Heward for comments on an early draft of the manuscript.

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249 **Figure 1. Map of the study area, Tree Sparrow colonies are marked as black circles.**
250 **Groups of nest boxes that were separated by more than 400m were defined as colonies.**

251 **Table 1. Results of GLMMs for the effect of Grass ES and WBS on Tree sparrow**
 252 **nestling diet. Models were run using binomial error distributions. Each dietary group**
 253 **was modelled separately. The estimated slope (\pm SE), Wald test statistic (z-value) and p-**
 254 **value significance are given. Grass ES and WBS habitat coverage was arcsine square**
 255 **root transformed to normalise their distribution.**

Response	Explanatory	Estimate \pm SE	z-value	p
Araneae	Intercept	1.26 \pm 0.72	1.74	0.082
	Grass ES	-3.80 \pm 3.13	-1.21	0.225
	WBS	5.61 \pm 3.56	1.57	0.116
Carabidae	Intercept	1.93 \pm 0.71	2.71	<0.01
	Grass ES	-0.58 \pm 3.12	-0.19	0.851
	WBS	-2.10 \pm 3.14	-0.67	0.502
Other Coleoptera adults	Intercept	1.12 \pm 1.40	0.80	0.425
	Grass ES	3.01 \pm 4.67	0.65	0.519
	WBS	-1.69 \pm 8.12	-0.21	0.835
Coleoptera larvae	Intercept	-1.35 \pm 1.42	-0.95	0.342
	Grass ES	3.36 \pm 5.37	0.63	0.532
	WBS	2.13 \pm 5.79	0.37	0.713
Diptera	Intercept	3.24 \pm 0.97	3.35	<0.001
	Grass ES	-2.36 \pm 3.79	-0.62	0.534
	WBS	-3.76 \pm 4.00	-0.94	0.348
Lepidoptera larvae	Intercept	-1.06 \pm 1.18	-0.94	0.349
	Grass ES	0.53 \pm 4.47	0.12	0.906
	WBS	-0.39 \pm 5.18	-0.08	0.940
Tipulidae	Intercept	2.18 \pm 1.40	1.56	0.119
	Grass ES	-5.66 \pm 4.70	-1.20	0.229
	WBS	-4.18 \pm 5.86	-0.71	0.475
Seed	Intercept	-0.59 \pm 1.01	-0.59	0.557
	Grass ES	1.66 \pm 4.41	0.38	0.707
	WBS	12.80 \pm 5.67	2.26	<0.05

256