The effect of silicon on iron plaque formation and arsenic accumulation in rice genotypes with different radial oxygen loss (ROL)

by Wu, C., Zou, Q, Xue, S., Pan, W., Huang, L., Hartley, W. Mo, J. and Wong, M.

Copyright, Publisher and Additional Information: This is the author accepted manuscript. The final published version (version of record) is available online via Elsevier Please refer to any applicable terms of use of the publisher.

DOI: http://dx.doi.org/10.1016/j.envpol.2016.01.004



Wu, C., Zou, Q, Xue, S., Pan, W., Huang, L., Hartley, W. Mo, J. and Wong, M. 2016. The effect of silicon on iron plaque formation and arsenic accumulation in rice genotypes with different radial oxygen loss (ROL). *Environmental Pollution*, 212, pp.27-33.

1	The effect of silicon on iron plaque formation and arsenic accumulation in rice
2	genotypes with different radial oxygen loss (ROL)
3	Chuan WU ¹ , Qi ZOU ¹ , Sheng-Guo XUE ^{1,*} , Wei-Song PAN ² , Liu HUANG ¹ , William
4	HARTLEY ³ , Jing-Yu MO ¹ Ming-Hung Wong ^{4,5}
5 6	1 School of Metallurgy and Environment, Central South University, Changsha 410083, China. E-mail:wuchuan@csu.edu.cn
7 8	2 College of Bioscience and Biotechnology, Hunan Agricultural University, Changsha 410128, China
9 10	3 Crop and Environment Sciences Department, Harper Adams University, Newport, Shropshire, TF10 8NB, United Kingdom
11 12	 4 Consortium on Health, Environment, Education and Research (CHEER), Hong Kong Institute of Education, Tai Po, Hong Kong Special Administrative Region 5 School of Environment, Linear University, Connection, Ching
13	5 School of Environment, Jinan University, Guangzhou, China
15	ABSTRACT Rice is one of the major exposure pathways of arsenic (As) in the human
16	food chain which threatens over half the global population. A series of greenhouse pot
17	investigations were conducted to examine the effects of Si application on iron plaque
18	formation, As uptake and rice grain As speciation in indica and hybrid rice genotypes
19	with different radial oxygen loss (ROL) patterns. The results demonstrated that Si
20	significantly increased root ($p < 0.05$) and grain ($p < 0.001$) biomass. Indica genotypes
21	with higher ROL induced greater Fe plaque formation compared to hybrid genotypes
22	(p<0.005) and sequestered more As in Fe plaque. Silicon applications significantly
23	increased Fe concentrations in iron plaque of different genotypes. Silicon application
24	significantly decreased As concentrations in roots ($p < 0.005$), straws ($p < 0.05$) and
25	husks ($p < 0.001$) by 28-35% $15-35\%$ and 32-57% respectively, whilst also reducing
26	inorganic As (iAs) and DMA in rice grains. Indica genotypes with higher ROL
27	accumulated lower concentrations of iAs in grains than hybrid genotypes with lower

28 ROL.

29 *Key words*: Arsenic, Iron plaque, Rice, Silicon.

30

49

31 **1. Introduction**

32 Arsenic (As) is a well-known carcinogenic metalloid and its exposure to humans is predominantly through drinking water and diet, which has led to increased health 33 risks (Stone, 2008; Cui et al., 2013). Paddy soils have been contaminated with As due 34 to the use of As-contaminated groundwater for irrigation, mining and other 35 36 anthropogenic activities (Rahman et al., 2014; Jia et al., 2014; Shi et al., 2014); this has resulted in the accumulation of soil As and hence its transfer into rice (Abedin and 37 Meharg, 2002; Seyfferthet et al., 2014). To exacerbate the problem, rice is largely 38 39 cultivated under flooded conditions, the anaerobic environment significantly enhancing the mobilization and bioavailability of As (Pan et al., 2014). Under 40 anaerobic conditions, arsenite is the predominant As species in soil solution 41 42 (Takahashi et al., 2004; Xu et al., 2008). Arsenic species in rice comprise mainly of 43 inorganic As (iAs), including arsenite and arsenate, monomethylarsonic acid (MMA) and dimethylarsinic acid (DMA) (Zhao et al., 2010; Nookabkaew et al., 2013). The 44 main As species in rice grain are iAs and DMA (Zhu et al., 2008a), with inorganic 45 species being considered of greater toxicity than MMA and DMA (Williams et al., 46 2005; EFSA, 2009; Calatayud et al., 2013), and thereby creating a major exposure 47 48 pathway via human ingestion (EFSA, 2009; Halder et al., 2012; Qu et al., 2015).

There are two main pathways for As uptake in rice (Zhao et al., 2013a). Firstly,

arsenate is a chemical analogue of phosphate, and arsenate can be assimilated in the 50 rice root via the phosphate transporter protein system (Chen et al., 2013; Wu et al., 51 52 2015). Secondly, arsenite is a silicic acid analogue, and therefore it can be taken up by roots through the silicic acid transport system (Ma et al., 2008; Chen et al., 2012). 53 54 Studies have also indicated that the addition of Si markedly decreased uptake and shoot As accumulation in rice (Guo et al., 2007; Wu et al., 2015). Furthermore, 55 Seyfferth and Fendorf (2012) discovered that the addition of Si in soil pore-water 56 significantly decreased As concentrations in rice grains. In addition, MMA also shares 57 58 the same silicic transportation pathway and it is reported that Si can promote As methylation and hence affect the concentration and species of As in rice grain. This 59 has been observed to reduce iAs concentrations by 59% whilst increasing DMA 60 61 concentrations by 33% (Li et al., 2009). In addition, it has been reported that Si application reduced As concentrations in straw, flag, leaf and husk by half, with 62 arsenite concentrations in brown and polished rice reduced by 22% and 33% 63 respectively (Fleck et al., 2013). It has also been demonstrated the Si application 64 strongly reduced the concentration of iAs, mainly arsenite, whilst increasing the 65 concentration of DMA in both vegetative and reproductive tissues of rice (Liu et al., 66 2014). Nanoscale silica sol foliar application may also alleviate toxicity and 67 accumulation of As in rice grains due to strengthening of their antioxidant defense 68 capacity (Liu et al., 2014). 69

The oxygenation of plant roots by radial oxygen loss ROL (Colmer, 2003) and rhizosphere oxygenation by microbial activities, converts Fe²⁺ to Fe³⁺, leading to the

formation of Fe-plaque around roots (Mei et al., 2009; Wu et al., 2012). Root ROL 72 rates are considered the key biotic factor controlling Fe-plaque formation (Li et al 73 74 2011; Wu et al., 2012) and the plaque is mainly comprised of ferric hydroxides (63%), goethite (32%) and minor concentrations of siderite (5%), whose structure is 75 characterized as amorphous or crystalline iron (oxyhydr)oxides (Liu et al., 2004a). 76 Due to adsorption or coprecipitation mechanisms, iron plaque can sequester metals, 77 metalloids (eg. As) and anions such as carbonate and silicate on rice roots (Liu et al., 78 2004a,b; Liu and Zhu, 2005). It has been widely demonstrated that iron plaque plays 79 80 an important role in mediating As accumulation and alleviating As toxicity in rice plants (Ultra et al., 2009; Wu et al., 2012; Lee et al., 2012); the iron plaque serving as 81 82 a barrier to prevent As translocation from roots to shoots (Liu et al., 2004a,b). Lee et 83 al., (2012) reported that As addition induced iron plaque formation on roots, the Fe formation decreasing As uptake by roots and shoots, indicating that iron plaque can 84 sequestrate As and reduce As uptake in rice. Furthermore, Wu et al (2012) 85 86 discovered that higher rates of ROL contributed to increases in Fe-plaque which subsequently sequestered more As on rice roots. Nevertheless, these previous studies 87 have focused on Fe plaque formation and Si concentration independently affecting As 88 accumulation, and showing limited evidence as to the effects of Si on Fe plaque 89 formation and As accumulation in rice genotypes with different ROL. 90

The aims of the present study were 1) to investigate the effects of Si on Fe plaque formation in rice genotypes with different ROL; 2) to determine the effect of Si on As sequestration in Fe plaque of rice genotypes with different ROL; 3) to determine the effect of Si on As concentrations and speciation in rice genotypes withdifferent ROL.

96

97 **2. Materials and methods**

98 2.1	Materials
--------	------------------

Rice seeds from four genotypes were obtained from Hunan Agricultural 99 University, and included hybrid subspecies Xiangfengyou 9 ('XFY-9'), T-you207 100 ('TY-207') and indica subspecies Xiangwanxian 17 ('XWX-17'), Xiangwanxian 12 101 102 ('XWX-12'). The ROL of XFY-9, TY-207, XWX-17 and XWX-12 were 9.55, 15.41, 19.76 and 27.00(需要单位吗) respectively (Wu et al., 2015). All seeds were surface 103 disinfected with a 30% H₂O₂ solution for 15 min, and subsequently washed repeatedly 104 105 in deionized water. The seeds were then germinated in culture dishes on moist filter paper. Germinated rice seedlings were then cultured in a nutrient solution for 2 weeks. 106

107

108 2.2 Pot investigation under waterlogged conditions

Paddy soils (pH 6.6, 9.4 mg As /kg) were collected (0-20 cm depth) from a paddy 109 field near the campus of Central South University, Changsha, P.R. China. Soils were 110 returned to the laboratory and air dried at room temperature, then ground and sieved 111 112 to <2 mm. Nutrients were then thoroughly mixed by hand into the soil as follows, P as CaH₂PO₄ H₂O at 0.15 g/kg P₂O₅, K as KCl at 0.2 g/kg K₂O, and N as CO (NH₂)₂ at 113 0.2 g/kg N (Wu et al., 2011). Arsenate solution (Na₂HAsO₄·12H₂O) was then applied 114 at 60 mg As/kg to all treatments with the exception of the control. Silicon was then 115 116 added as a SiO₂ colloid (63-200µm) (Seyfferth and Fendorf, 2012) as follows: Control, no Si, no As; Treatment A, arsenate only (Si0); Treatment B, arsenate and 10 mg Si/kg 117 (Si10); Treatment C, arsenate and 20 mg Si/kg (Si20); Treatment D, arsenate and 40 118

mg Si/kg (Si40). All treatments were subsequently thoroughly hand-mixed and 119 allowed to equilibrate for two weeks. After equilibration, seedlings (three seedlings 120 per pot) from the four rice genotypes were transferred to polyethylene pots (20 cm 121 diameter, 20 cm high) which had been previously filled with 3.5 kg of the individual 122 soil treatments. All treatments were carried out in triplicate. After transplantation, the 123 124 seedlings were grown under waterlogged conditions, the water level being maintained at 2-3 cm above the soil surface. The pots were placed randomly in a greenhouse 125 (25°C day, 20°C night, relative humidity 70%) and natural light was supplemented 126 with sodium light (1200 Lux), providing a photoperiod of 12 hr light/12 hr dark. 127 Plants were harvested after maturity. 128

129

130 *2.3 Extraction of Fe plaque*

Rice plants were harvested at maturity and washed thoroughly using tap water 131 followed by deionized water. Plants were then divided into roots and shoots. Iron 132 133 plaque on root surfaces was determined by DCB-extraction (Wu et al., 2012). Root tissue (1.0 g) was extracted using 30 ml of DCB solution for 30 min at room 134 temperature; the solution prepared using 0.03 Μ sodium citrate 135 was 136 (Na₃C₆H₅O₇.2H₂O) and 0.125 M sodium bicarbonate (NaHCO₃), with 0.6 g of sodium dithiomite (Na₂S₂O₄). After incubation, roots were rinsed three times with deionized 137 water and the washings were added to the DCB extract. Deionized water was then 138 added to the extracting solution to obtain 100 ml solution prior to analysis. The Fe 139 concentrations in extracts were measured by Atomic Absorption Spectrometry (AAS, 140 TAS-990, Beijing Puxi Instruments Co., P.R. China). Arsenic concentrations in 141 142 extracts were determined with hydride generation atomic fluorescence spectrometry (HG-AFS, AFS-8230, Beijing Jitian Instruments Co., China). 143

145 2.4 Plant analysis for total As

Plant samples were harvested at maturity and washed carefully using deionized water and divided into root, straw, husk and grain. Straw from below the irrigation water was removed to avoid contamination. Half the roots, straw and grains were oven-dried at 70°C to a constant weight, while the other half was freeze dried for As speciation determination. Grains and husks were further divided, ground using a pestle and mortar, and stored at -20°C prior to analysis.

152 After dry weight determination, samples were ground using a mechanical mill, and 0.5g sample was weighed into a conical flask (100ml) with 5ml concentrated 153 nitric acid. The samples were left to digest overnight at room temperature, then placed 154 155 on an electric hot plate (120°C) until the solution became clear. After digestion, the samples were filtered (0.45 µm) into colorimetric tubes and diluted to 20 ml with 156 deionized water. The total As concentration (root, straw, husk and grain) was 157 determined using HG-AFS (AFS-8230, Beijing Jitian Instruments Co., China) (Zhu et 158 al., 2008b). A certified reference material (bush branches and leaves, GBW07603) 159 was used for quality control purposes, with recovery of As ranging from 80.4% to 160 89.5%. 161

162

163 2.5 Plant analysis for As speciation

For determination of As species in grain, samples were ground under liquid N₂ to
ensure stabilization of As species (Zhu et al., 2008b). Milled grain samples (1.0 g)

166	were added to centrifuge tubes (50 ml), and 20 ml HNO ₃ (1%) was added and heated
167	to 95°C for 1.5 h. After the samples had cooled to room temperature, the extracting
168	solution was centrifuged at 5000r/min for 10 min and the supernatant filtered (0.22
169	μ m). Arsenic speciation was determined using HPLC-HG-AFS (HPLC, Shimadzu
170	LC-15C Suzhou Instruments Co., China; HG-AFS, AFS-8230, Beijing Jitian
171	Instruments Co., China) (Zhu et al., 2008b; Shi et al., 2013).

173	2.6 Data analysis
-----	-------------------

All data was analyzed in EXCEL 2007. Analysis of variance for plant biomass,
As and Fe concentrations in different genotypes and Si treatments were determined by
SPSS 19.0. All figures were created in Origin 8.0.

177

178 **3. Results**

179 *3.1 Effect of Si on plant growth*

Root, straw and grain biomass, from the four rice genotypes grown in different 180 treatments, is presented in table 1. Significant differences are observed in root 181 biomass between genotypes (p < 0.05) (Table 1). In the control treatment, root 182 biomass was ranked as XWX-12 < TY-207 < XWX-17 < XFY-9, ranging from 9.25 183 g/pot to 15.8 g/pot. Under different Si concentrations, XFY-9 and XWX-12 developed 184 the largest root biomass, whilst TY-207 and XWX-17 displayed the greatest root 185 biomass in Si20 treatment (Table 1). Silicon application significantly increased root 186 biomass of XFY-9, XWX-12 and TY-207 (p<0.05), but had no significant effect on 187

188 XWX-17 (p>0.05). Root biomass of XFY-9 was significantly greater than the other 189 three genotypes in Si40 treatment (p<0.05) (Table 1).

190 Significant genotypic differences were also observed in straw biomass (p < 0.001). In control treatments, straw biomass from the four genotypes ranged from 17.8 g/pot to 191 29.4 g/pot (XWX-12 < TY-207 < XFY-9 < XWX-17). Straw biomass from XWX-17 192 was the greatest between the four genotypes in Si40 treatment (p < 0.05) (Table 1). 193 Results showed that there were significant differences in grain biomass between 194 genotypes (p < 0.001). In control treatments, grain biomass of the four genotypes was 195 ranked as follows, XWX-12 < TY-207 < XFY-9 < XWX-17, with the largest value 196 being 14.8 g/pot and 6.4 g/pot. Silicon significantly increased grain biomass of the 197 four genotypes (p<0.001) (Table 1). Genotypes XFY-9, XWX-17 and XWX-12 198 199 obtained the largest grain biomass within Si40 treatment, with values being 17.0, 19.1 and 12.9 g/pot respectively; the grain biomass of TY-207 was greatest in Si-20 200 treatment (18.2 g/pot). 201

202

203 *3.2 Effect of Si on Fe plaque formation*

In control treatments, As concentrations in Fe plaque were below the detection limit. Significant differences (p<0.005) in Fe concentrations in Fe plaque were observed between the genotypes. In control treatments, Fe concentrations of the four genotypes ranked as follows, XFY-9 < XWX-17 < TY-207 < XWX-12. The results demonstrated that Si had a significant effect on Fe concentrations in plaque formation (p<0.05). Silicon additions increased Fe concentrations in Fe plaque (Figure 1a) and had a significant effect on As concentrations in the Fe plaque. However, withincreasing Si, As concentrations firstly increased but then decreased (Figure 1b).

212 Both Fe and As concentrations were higher in DCB-extracts from indica than that of the other hybrid genotypes (Figure 2), demonstrating that indica genotypes 213 with higher ROL contribute to further Fe plaque formation. Compared to treatments 214 without Si, Si application influenced the sequestration of As on Fe plaque both in 215 indica and hybrid genotypes (Figure 2a, b). Arsenic was positively correlated with 216 Fe concentrations in Si0 treatment (Figure 2a), however, with the increasing Fe 217 218 concentrations, As concentrations increased only marginally in hybrid genotypes, and 219 negatively correlated with Fe concentrations in indica genotypes in Si treatments 220 (Figure 2b).

221

222 3.3 The effect of Si on total As in rice plants

Total As concentrations in rice plants cultivated in control treatments were below the 223 detection limit (Table 3). There were significant genotypic effects on root As 224 concentrations in rice (p < 0.05) (Figure 3a). Application of Si significantly reduced 225 root As concentrations in rice plants (p < 0.005), and with increasing Si concentrations, 226 root As concentrations in the four genotypes decreased, reaching the lowest 227 concentrations in Si40 treatment (Figure 3a). In Si40 treatment, root As 228 concentrations from the four genotypes were reduced to their lowest. Compared to Si0 229 treatment, root As concentrations decreased 30%, 35%, 28% and 16% in the four 230 treatments respectively. 231

Genotypes had no significant effect on straw As concentrations in rice plants 232 (p>0.05) (Figure 3b). Genotypes XWX-17 and XWX-12 grown in Si40 treatment 233 however differed significantly in husk total As concentrations (p < 0.01) (Figure 3c). 234 Genotypes had a significant effect on total grain As concentrations (p < 0.001) 235 (Figure 3d). In Si0 treatment, As concentrations in grains ranked as follows, XWX-12 236 < XWX-17 < TY-207 < XFY-9. A significant effect on reducing grain As 237 concentrations in XWX-17 grown in Si10, Si20 and Si40 treatments was also 238 observed compared to Si0 treatment (Figure 3d). 239

240

241 *3.4 The effects of Si on As species*

The predominant As species in rice grain were As (III) and DMA; DMA accounted for 91%-95% of total As concentrations in grains (Figure 4). All Si treatments reduced DMA accumulation in rice grains of genotype XWX-12 (Figure 4). The Si40 treatment decreased total grain As of XWX-12 by 20% compared to Si0 treatment. Results also demonstrated that MMA was not detectable in the grain. Silicon addition decreased DMA accumulation but had no significant effect on iAs accumulation.

249

250 **4. Discussion**

251 4.1 Effect of Si on rice biomass

252 Rice is a typical Si-accumulating crop, and Si application, regardless of Si 253 solution, minerals or foliar-fertilizer, is considered to have positive and consistent

effects on rice health and yield (Seyfferth and Fendorf, 2012; Detmann et al., 2012; Liu 254 et al., 2014), due to alleviation of both biotic and abiotic stress factors (Ma, 2004; 255 256 Guo et al., 2005; Epstein, 2009). Emerging evidence has shown that Si can enhance rice resistance to As toxicity (Fleck et al., 2013). Silicon increased shoot length and 257 biomass of rice seedlings cultivated in As-amended hydroponic solutions (Guo et al., 258 2007; Guo et al., 2009; Tripathi et al., 2013) and this has been demonstrated in our pot 259 experiment indicating that Si significantly increased the biomass of roots and grains 260 of different rice genotypes (Table 1). Recent studies have also reported that Si 261 262 markedly increased rice grain and straw yield (Li et al., 2009; Fleck et al., 2013; Liu et al., 2014), which is consistent with our results, but we also observed that with 263 certain Si treatments the opposite effect can occur (Table 1). Recent research by Lee 264 265 et al (2014) demonstrated that Si addition to As-contaminated soil increased As concentrations in soil solution due to competition between of Si and As for adsorption 266 sites; this increased the bioavailability of As and increased its toxicity to rice. 267

268

269 4.2 Effect of Si on Fe plaque formation

Iron plaque is abundantly formed on the root surfaces of common aquatic plants such as paddy rice (Chen et al., 2005; Wu et al., 2012). The formation of the plaque is considered to be a "barrier" to As uptake and subsequent above-ground accumulation in rice (Liu et al., 2004a; Wu et al., 2012). Previous studies have demonstrated that there are many factors influencing the formation of Fe plaque such as radial oxygen loss (ROL), genotype, pH, Eh and presence of microorganisms (Taylor et al., 1984;

Emerson et al., 1999; Wu et al., 2012). Iron oxides are able to combine with silicate to 276 form iron silicate and deposit on root surfaces (Liu et al., 2004a, b; Liu and Zhu, 277 2005). Results from the present study have indicated that Si application can 278 significantly increase Fe plaque formation (p < 0.05). Sevfferth and Fendorf (2012) 279 discovered that Si concentrations, in root Fe plaque of three rice genotypes, were 280 higher in Si treatments than controls (no Si), which is further confirmed in our 281 investigation. Liu et al (2006) observed that Fe plaque mainly consisted of oxides, but 282 also contained ferrihydrite and goethite. Silicon acid limits the adsorption of As(III) 283 284 and As(V) on ferrihydrite by occupying the absorption sites (Luxton et al., 2008; Liu et al., 2014) and this may explain the decrease in As observed in DCB-extracts with 285 increasing Si concentrations (Figure 2b). 286

287

288 *4.3 Effect of Si on As uptake by rice*

Previous studies have revealed the suppressing effect of Si on As uptake and 289 290 accumulation in rice, regardless of whether the investigation was hydroponic (Guo et al., 2005, 2007) or pot experiments (Seyfferth and Fendorf, 2012; Fleck et al., 2013; 291 Wu et al., 2015). Numerous studies have indicated that As(III) is the predominant 292 species in waterlogged soils due to the absence of O_2 (Takahashi et al., 2004; Xu et al., 293 2008; Zhao et al., 2013a) and As(III) concentrations can exceed 87% of the total As in 294 soil solution (Liu et al., 2014). As a chemical analogue of silicic acid, As(III) is 295 taken up into rice roots via the Si transport channel (Ma et al., 2008). 296 This consequently leads to competition between Si and arsenite for uptake and 297

translocation in rice roots. However, Si application has been shown to reduce the 298 regulation of Si transporter genes (Ma et al., 2006; Liu et al., 2014) and Fleck et al 299 300 (2013) reported that Si application decreased As concentrations in brown and polished rice by 22% by inhibiting As(III) uptake and transportation to shoots. Furthermore, 301 external Si increased the P concentration in nutrient solution and soil pore water in 302 hydroponic and soil pot experiments respectively (Guo et al., 2007; Fleck et al., 2013). 303 In addition, while As(V) has similar physicochemical properties with phosphate and 304 As(V) shares the same phosphate transport pathway (Abedin et al., 2002), increased P 305 306 may inhibit As uptake in rice. Although some studies have reported that Si addition increased As concentrations in soil solution, by displacing soil adsorption retention 307 sites and inhibiting As adsorption on Fe plaque (Luxton et al., 2008; Liu et al., 2014), 308 309 the effects were much lower than the inhibitory effect of Si on As uptake and translocation in rice roots (Liu et al., 2014). The competitive relationship between Si 310 and As was determined by Lee et al (2014), who observed that higher Si/As ratios 311 312 induced greater inhibitory effects on As uptake.

313

314

315 *4.4 Effect of Si on As speciation in rice grain*

Previous investigations have demonstrated that iAs and DMA were the predominant As species in rice grain (Zavala et al., 2008; Zhu et al., 2008a), which are consistent with the results of the present study (Figure 4). Not only did the addition of Si reduce total As concentrations in XFY-9 and XWX-12, but it reduced iAs by 16

and 20% respectively (Table 4), this result being consistent with Li et al (2009) who 320 observed a reduction iAs by 59% following Si application. Arsenite is the main As 321 322 species found in rhizospheres of waterlogged soils (Takahashi et al., 2004; Xu et al., 2008; Zhao et al., 2013a). As a result of competition between Si and As(III) in the 323 silicic acid transport system, and the reduced expression of the Si transporter genes 324 (Ma et al., 2006; Liu et al., 2014), Si reduced the uptake and translocation of As in 325 rice plants (Liu et al., 2014). Furthermore, rice grain iAs concentrations were lower 326 in genotypes with higher ROL than genotypes with lower ROL under the same Si 327 328 treatments. This indicates that both genotype and environmental conditions have an influence on As speciation. 329

In the present study Si reduced DMA concentrations in rice grain, whereas Li et al 330 331 (2009) and Liu et al (2014) observed that Si application increased DMA concentrations in rice grain. Conflicting evidence from Fleck et al (2013) reported 332 that Si application to soil didn't have any influence on DMA concentrations in rice 333 334 grains. Recent research indicates that rice plants lack the ability to methylate As, and DMA is derived from methylation by soil microorganisms (Zhao et al., 2013a); 335 differences in microbial communities and soil type will affect this process 336 considerably (Zhao et al., 2013b). Additionally, Fleck et al (2013) demonstrated that 337 Si addition increased soil pH, and Li et al (2009) found that increasing soil pH 338 reduced MMA and DMA uptake significantly in rice. Additionally, in paddy soils it is 339 340 believed that DMA can exist as neutral molecules and negatively charged anions. Former species of DMA are most likely to be regulated by the silicic acid transport 341

system (Kersten and Daus, 2014), therefore external Si additions may inhibit the
uptake of DMA. Consequently, Si addition may have decreased DMA concentrations,
whilst preferential uptake of As(III) by roots may have further reduced DMA uptake
(Liu et al., 2014) as the affinity for As(III) is stronger than DMA (Abedin et al.,
2002).

347

348 **5.** Conclusion

Four rice genotypes which differed in their radial oxygen loss (ROL), two hybrid 349 350 subspecies Xiangfengyou 9 ('XFY-9') and T-you 207 ('TY-207'), and two indica subspecies Xiangwanxian 17 ('XWX-17') and Xiangwanxian 12 ('XWX-12'), 351 were selected to investigate the effects of varying silicon concentrations on As uptake, 352 353 iron plaque formation in rice and As speciation in grains. Results demonstrated that in most treatments, Si additions increased grain biomass. There were significant 354 genotypic differences in the concentrations of iron plaque formed, with indica 355 producing more iron plaque than hybrid genotypes and sequestering higher As 356 concentrations. Silicon applications significantly increased Fe concentrations in iron 357 plaque of the genotypes. Arsenic concentrations in roots, husks and grains of indica 358 genotypes were lower than hybrid genotypes. The main As species in rice grains were 359 As(III) and DMA, with DMA accounting for > 90% of total As. Indica genotypes 360 accumulated lower concentrations of iAs than hybrid genotypes. Silicon addition 361 decreased both inorganic arsenic and DMA concentrations in grain. This study is 362 potentially a step forward to understanding As uptake mechanisms in rice and 363

364 mitigating the health risks posed by As contamination in paddy fields.

365 Acknowledgment

- 366 Financial support from National Natural Science Foundation of China (No. 41201493;
- 367 31300815) and Natural Science Foundation of Hunan, China (No. 2015JJ3142) is
- 368 gratefully acknowledged. Chuan Wu acknowledged the Postdoctoral Science
- 369 Foundation of Central South University in Changsha, P.R. China. And thank you

370 Yang Fei, Zhaocai Li in Central South University for the help of experiments.

371

372 **References**

- Abedin, M. J., Meharg, A. A., 2002. Relative toxicity of arsenite and arsenate on
 germination and early seedling growth of rice (Oryza sativa L.). Plant. Soil. 243 (1),
 57–66.
- Abedin, M.J., Feldmann, J., Meharg, A.A., 2002. Uptake kinetics of arsenic species in
 rice plants. Plant Physiol. 128 (3), 1120–1128.
- Calatayud, M., Bralatei, E., Feldmann, J., Devesa, V., Vélez, D., 2013.
 Transformation of arsenic species during in vitro gastrointestinal digestion of
 vegetables. J. Agr. Food Chem. 61, 12164-12170.
- 381 Chen, X.W., Li, H., Chan, W.F., Wu, C., Wu, F.Y., Wu, S.C., Wong, M.H., 2012.
- 382 Arsenite transporters expression in rice (Oryza sativa L.) associated with arbuscular
- 383 mycorrhizal fungi (AMF) colonization under different levels of arsenite stress.
- 384 Chemosphere 89, 1248–1254.
- 385 Chen, X.W., Wu, F.Y., Li, H., Chan, W.F., Wu, C., Wu, S.C., Wong, M.H., 2013.
- 386 Phosphate transporters expression in rice (Oryza sativa L.) associated with arbuscular
- 387 mycorrhizal fungi (AMF) colonization under different levels of arsenate stress.
- 388 Environ. Exp. Bot. 87, 92-99.
- Chen, Z., Zhu, Y.G., Liu, W.j., Meharg, A.A., 2005. Direct evidence showing the
- 390 effect of root surface iron plaque on arsenite and arsenate uptake into rice (Oryza

- 391 sativa) roots. New. Phytol. 165, 91–97.
- Colmer, T.D., 2003. Long-distance transport of gases in plants: a perspective on
 internal aeration and radial oxygen loss from roots. Plant. Cell. Environ. 26 (1),
 17–36.
- Cui, J., Shi, J., Jiang, G., Jing, C., 2013. Arsenic levels and speciation from ingestion
- exposures to biomarkers in shanxi, china: Implications for human health. Environ. Sci.Technol. 47, 5419-5424.
- Detmann, K.C., Araújo, W.L., Martins, S.C.V., Sanglard, L.M.V.P., Reis, J.V.,
 Detmann, E., Rodrigues, F.A., Nunes-Nesi, A., Fernie, A.R., DaMatta, F.M., 2012.
 Silicon nutrition increases grain yield, which, in turn, exerts a feed-forward
 stimulation of photosynthetic rates via enhanced mesophyll conductance and alters
- 402 primary metabolism in rice. New. Phytol. 196, 752-762.
- 403 Emerson, D., Weiss, J.V., Megonigal, J.P., 1999. Iron-oxidizing bacteria are associated
- with ferric hydroxide precipitates (Fe-plaque) on the roots of wetland plants. Appl.
 Environ. Microb. 65(6), 2758-2761.
- 406 Epstein, E., 2009. Silicon: its manifold roles in plants. Ann. Appl. Biol. 155, 155–160.
- 407 European Food Safety Authority (EFSA)., 2009. Scientific opinion on arsenic in food.
- 408 EFSA Panel on Contaminants in the Food Chain (CONTAM). EFSA. J 7, 1531–1550.
- 409 Fleck, A.T., Mattusch, J., Schenk, M.K., 2013. Silicon decreases the arsenic level in
- rice grain by limiting arsenite transport. J. Plant. Nutr. Soil. Sci. 176, 785–794.
- Guo, W., Hou, Y.L., Wang, S.G., Zhu, Y.G., 2005. Effect of silicate on the growth
 and arsenate uptake by rice (Oryza sativa L.) seedlings in solution culture. Plant Soil
 272, 173-181.
- Guo, W., Zhang, J., Teng, M., Wang, L.H., 2009. Asenic uptake is suppressed in a rice
 mutant defective in silicon uptake. J. Plant Nutr. Soil Sci. 172, 867–874.
- 416 Guo, W., Zhu, Y.G., Liu, W.J., Liang, Y.C., Geng, C.N., Wang, S.G., 2007. Is the
- 417 effect of silicon on rice uptake of arsenate (AsV) related to internal silicon
- 418 concentrations, iron plaque and phosphate nutrition? Environ. Pollut. 148, 251-257.
- 419 Halder, D., Bhowmick, S., Biswas, A., Chatterjee, D., Nriagu, J., Guha Mazumder,
- 420 D.N., Šlejkovec, Z., Jacks, G., Bhattacharya, P., 2013. Risk of arsenic exposure from

- drinking water and dietary components: Implications for risk management in ruralbengal. Environ. Sci. Technol. 47, 1120-1127.
- Jia, Y., Huang, H., Chen, Z., Zhu, Y., 2014. Arsenic uptake by rice is influenced by
 Microbe-Mediated arsenic redox changes in the rhizosphere. Environ. Sci. Technol.
 48, 1001-1007.
- 426 Kersten, M., Daus, B., 2015. Silicic acid competes for dimethylarsinic acid (DMA)
- 427 immobilization by the iron hydroxide plaque mineral goethite. Sci. Total Environ. 508,428 199-205.
- 429 Lee, C.H., Hsieh, Y.C., Lin, T.H., Lee, D.Y, 2013. Iron plaque formation and its
- effect on arsenic uptake by different genotypes of paddy rice. Plant Soil 363, 231-241.
- 431 Lee, C.H., Huang, H.H., Syu, C.H., Lin, T.H., Lee, D.Y, 2014. Increase of As release
- and phytotoxicity to rice seedlings in As-contaminated paddy soils by Si fertilizer
- 433 application. J. Hazard. Mater. 276, 253-261.
- Li, H., Ye, Z.H., Wei, Z.J., Wong, M.H., 2011. Root porosity and radial oxygen loss related to arsenic tolerance and uptake in wetland plants. Environ. Pollut. 159, 30-37.
- Li, R.Y., Ago, Y., Liu, W.J., Mitani, N., Feldmann, J., McGrath, S.P., Ma, J.F., Zhao,
- F.J., 2009. The rice aquaporin Lsi1 mediates uptake of methylated arsenic species.
 Plant Physiol. 150, 2071–2080.
- Li, R.Y., Stroud, J.L., Ma, J.F., McGrath, S.P., Zhao, F.J., 2009. Mitigation of arsenic
- accumulation in rice with water management and silicon fertilization. Environ. Sci.Technol. 43, 3778-3783.
- Liu, C.P., Wei, L., Zhang, S.R., Xu, X.H., Li, F.B., 2014. Effects of nanoscale silica
- sol foliar application on arsenic uptake, distribution and oxidative damage defense in
 rice (Oryza sativa L.) under arsenic stress, RSC. Adv. 4, 57227.
- Liu, W.J., McGrath, S.P., Zhao, F.j., 2014. Silicon has opposite effects on the accumulation of inorganic and methylated arsenic species in rice. Plant. Soil. 376, 423-431.
- Liu, W.J., Zhu, Y.G. 2005. Iron and Mn plaques on the surface of roots of wetland
- 449 plants. Acta. Ecol. Sin, 25(2), 358-363
- Liu, W.J., Zhu, Y.G., Smith, F. A., Smith, S. E., 2004a. Do iron plaque and genotypes

- affect arsenate uptake and translocation by rice seedlings (Oryza sativa L.) grown in
 solution culture? J. Exp. Bot. 55, 1707-1713.
- Liu, W.j., Zhu, Y.G., Smith, F. A., Smith, S. E., 2004b. Do phosphorus nutrition and
- 454 iron plaque alter arsenate (as) uptake by rice seedlings in hydroponic culture. New.
- 455 Phytol. 162, 481–488.
- 456 Luxton, T.P., Eick, M.J., Rimstidt, D.J., 2008. The role of silicate in the
- adsorption/desorption of arsenite on goethite. Chem. Geol. 252(3), 125-135.
- Ma, J.F., 2004. Role of silicon in enhancing the resistance of plants to biotic and
 abiotic stresses. Soil Sci. Plant Nutr. 50 (1), 11–18.
- 460 Ma, J.F., Tamai, K., Yamaji, N., Mitani, N., Konishi, S., Katsuhara, M., Ishiguro, M.,
- 461 Murata, Y., Yano, M., 2006. A silicon transporter in rice. Nature. 440, 688–691.
- 462 Ma, J.F., Yamaji, N., Mitani, N., Xu, X.Y., Su, Y.H., McGrath, S.P., Zhao, F.J.,
- 463 2008. Transporters of arsenite in rice and their role in arsenic accumulation in rice
- 464 grain. Proc. Natl. Acad. Sci. USA 105 (29), 9931–9935.
- Mei, X.Q., Ye, Z.H., Wong, M.H., 2009. The relationship of root porosity and radial
 oxygen loss on arsenic tolerance and uptake in rice grains and straw. Environ. Pollut.
 157, 2550-2557.
- 468 Nookabkaew, S., Rangkadilok, N., Mahidol, C., Promsuk, G., Satayavivad, J., 2013.
- 469 Determination of arsenic species in rice from thailand and other asian countries using
- simple extraction and HPLC-ICP-MS analysis. J. Agr. Food Chem. 61, 6991-6998.
- 471 Pan, W.S., Wu, C., Xue S.G., Hartley, W., 2014. Arsenic dynamics in the rhizosphere
 472 and its sequestration on rice roots as affected by root oxidation. J. Environ. Sci. 26,
 473 892-899.
- 474 Qu, H., Mudalige, T.K., Linder, S.W., 2015. Arsenic speciation in rice by capillary
- 475 Electrophoresis/Inductively coupled plasma mass spectrometry: Enzyme-Assisted
- 476 Water-Phase microwave digestion. J. Agr. Food Chem. 63, 3153-3160.
- 477 Rahman, M.A., Rahman, M.M., Reichman, S.M., Lim, R.P., Naidu, R., 2014. Arsenic
- 478 speciation in Australian-Grown and imported rice on sale in australia: Implications for
- human health risk. J. Agr. Food Chem. 62, 6016-6024.
- 480 Seyfferth, A.L., Fendorf, S., 2012. Silicate Mineral Impacts on the Uptake and

- 481 Storage of Arsenic and Plant Nutrients in Rice (Oryza sativa L.). Environ. Sci.
 482 Technol. 46, 13176-13183.
- Seyfferth, A.L., McCurdy, S., Schaefer, M.V., Fendorf, S., 2014. Arsenic
 concentrations in paddy soil and rice and health implications for major Rice-Growing
 regions of Cambodia. Environ. Sci. Technol. 48, 4699-4706.
- 486 Shi, G.L., Lou, L.Q., Zhuang, S., Xia, X.W., Cai, Q.S., 2013. Arsenic, copper and
- 487 zinccontamination in soil and wheat during coal mining with assessment of health
- risks for the inhabitants of Huaibei, China. Environ. Sci. Pollut. Res. 20 (12),
 8435–8445.
- Shi, T., Liu, H., Wang, J., Chen, Y., Fei, T., Wu, G., 2014. Monitoring arsenic
 contamination in agricultural soils with reflectance spectroscopy of rice plants.
 Environ. Sci. Technol. 48, 6264-6272.
- 493 Stone, R., 2008. Arsenic and paddy rice: a neglected cancer risk. Nature 321,
 494 184–185.
- Takahashi, Y., Minamikawa, R., Hattori, K.H., Kurishima, K., Kihou, N., Yuita, K.,
 2004. Arsenic behavior in paddy fields during the cycle of flooded and non-flooded
 periods. Environ. Sci. Technol. 38, 1038-1044.
- Taylor, G.J., Crowder, A.A., Rodden, R., 1984. Formation and morphology of an iron
- plaque on the roots of Typha latifolia L. grown in solution culture. Amer. J. Bot. 71(5),666-675.
- 501 Tripathi, P., Tripathi, R.D., Singh, R.P., Dwivedi, S., Goutam, D., Shri, M., Trivedi,
- 502 P.K., Chakrabarty, D., 2013. Silicon mediates arsenic tolerance in rice (Oryza sativa
- 503 L.) through lowering of arsenic uptake and improved antioxidant defence system.
- 504 Ecol. Eng. 52, 96-103.
- 505 Ultra Jr, V.U., Nakayama, A., Tanaka, S., Kang, Y., Sakurai, K., Iwasaki, K., 2009.
- 506 Potential for the alleviation of arsenic toxicity in paddy rice using amorphous 507 iron-(hydr) oxide amendments[J]. Soil. Sci. Plant. Nutr. 55(1), 160-169.
- 508 Williams, P.N., Price, A.H., Raab, A., Hossain, S.A., Feldmann, J., Meharg, A.A.,
- 509 2005. Variation in arsenic speciation and concentration in paddy rice related to dietary
- 510 exposure. Environ. Sci. Technol. 39, 5531-5540.

- 511 Wu, C., Ye, Z., Li, H., Wu, S., Deng, D., Zhu, Y., Wong, M., 2012. Do radial oxygen
- 512 loss and external aeration affect iron plaque formation and arsenic accumulation and
- 513 speciation in rice? J. Exp. Bot. 63, 2961-2970.
- 514 Wu, C., Ye, Z.H., Shu, W.S., Zhu, Y.G, Wong, M.H., 2011. Arsenic accumulation
- and speciation in rice are affected by root aeration and variation of genotypes. J. Exp.Bot. 62, 2889-2898.
- 517 Wu, C., Zou, Q., Xue, S., Mo, J., Pan, W., Lou, L., Wong, M.H., 2015. Effects of
- silicon (Si) on arsenic (as) accumulation and speciation in rice (Oryza sativa L.)
- 519 genotypes with different radial oxygen loss (ROL). Chemosphere 138, 447-453.
- 520 Xu, X.Y., McGrath, S.P., Meharg, A.A., Zhao, F.J., 2008. Growing rice aerobically
- 521 markedly decreases arsenic accumulation. Environ. Sci. Technol. 42, 5574-5579.
- 522 Zavala, Y.J., Gerads, R., Gürleyük, H., Duxbury, J.M., 2008. Arsenic in rice: II.
- Arsenic speciation in USA grain and implications for human health. Environ. Sci.
 Technol. 42, 3861-3866.
- Zhao, F.J., McGrath, S.P., Meharg, A.A., 2010. Arsenic as a food chain contaminant:
 Mechanisms of plant uptake and metabolism and mitigation strategies. Annu. Rev.
 Plant Biol. 61, 535-559.
- Zhao, F.J., Zhu, Y.G, Meharg, A.A., 2013a. Methylated arsenic species in rice:
 Geographical variation, origin, and uptake mechanisms. Environ. Sci. Technol. 47,
 3957-3966.
- Zhao, F.J., Harris, E., Yan, J., Ma, J.C., Wu, L.Y., Liu, W.J., McGrath, S.P., Zhou,
 J.Z., Zhu, Y.G., 2013b. Arsenic methylation in soils and its relationship with
 microbial arsM abundance and diversity, and As speciation in rice. Environ. Sci.
 Technol. 47, 7147–7154.
- 535 Zhu, Y G., Williams, P.N., Meharg, A.A., 2008a. Exposure to inorganic arsenic from
- rice: A global health issue? Environ. Pollut. 154, 169-171.
- 537 Zhu, Y.G., Sun, G.X., Lei, M., Teng, M., Liu, Y.X., Chen, N.C., Wang, L.H., Carey,
- A.M., Deacon, C., Raab, A., Meharg, A.A., Williams, P.N., 2008. High percentage
- 539 inorganic arsenic content of mining impacted and nonimpacted Chinese rice. Environ.
- 540 Sci. Technol. 42 (13), 5008–5013.

Genotype	Treatment	Root	Straw	Grain
XFY-9	Control	15.8±2.73	22.0±4.34	7.80±0.64
	Si0	16.9 ± 1.50	24.1±3.78	10.6±0.80
	Si10	14.1±1.69	21.8±2.06	4.43±1.78
	Si20	15.7 ± 5.64	29.4±1.63	14.3±0.92
	Si40	25.1±2.60	26.3±3.62	17.0±0.7
TY-207	Control	9.7±0.11	19.7 ± 0.68	11.2 ± 1.7
	Si0	17.5 ± 2.52	22.6±6.26	11.5±1.0
	Si10	18.8 ± 7.06	24.2 ± 6.82	9.19±1.6
	Si20	22.3±2.04	28.8 ± 5.75	18.2±2.4
	Si40	17.9 ± 3.14	26.5±3.44	15.5±1.7
XWX-17	Control	11.5 ± 2.06	29.4±1.30	14.8±2.8
	Si0	17.6 ± 8.82	27.1±12.0	12.6±2.0
	Si10	14.0 ± 4.28	24.9 ± 5.96	14.5±4.12
	Si20	17.8±6.13	32.6±9.59	10.5±0.9
	Si40	14.6 ± 2.97	35.6±4.68	19.1±4.9
XWX-12	Control	9.25±0.94	17.8 ± 3.18	$6.40{\pm}1.8$
	Si0	12.4 ± 2.43	18.1±4.80	8.47 ± 1.1
	Si10	12.1±4.19	15.8 ± 6.00	11.3±1.0
	Si20	$12.7{\pm}1.85$	16.2 ± 7.01	12.1±0.7
	Si40	13.2±1.63	18.3±3.59	12.9±0.9
Analysis of variar	nce			
Genotype (G)		P<0.05	P<0.001	P<0.001
Si		P<0.05	NS	P<0.001
G×Si		NS*	NS	P<0.001

Table 1. Biomass of roots, straws and grains from the four genotypes grown in the

Genotypes	Treatments	Fe (mg/kg)	As (mg/kg)
XFY-9	Control	616±75.2	ND ^a
	SiO	835±139	20.8±7.5
	Si 10	1119±160	24.4±10.2
	Si20	1240±298	21.5±16.9
	Si40	1160±158	19.1±6.4
TY-207	Control	925±141	ND
	Si0	900±154	21.3±9.7
	Si 10	1080±248	26.6±12.7
	Si20	881±86.9	21.8±7.2
	Si40	962±301	18.4±5.7
XWX-17	Control	873±307	ND
	Si0	1130±707	36.9 ±28.1
	Si10	1430±603	40.5±20.8
	Si20	1210±626	32.1±22.7
	Si40	1380±191	25.4±5.8
XWX-12	Control	985±118	ND
	Si0	1450±44.2	25.5±5.6
	Si10	1490±593	31.3±15.0
	Si20	1500±255	24.1±7.1
	Si40	1660±49.8	27.1±4.5
Analysis	of variance		
Genotype(G)		P<0.005	NS
Si		P<0.05	P<0.001
G	< S i	NS^b	NS

Table 2. Iron and As concentrations in Fe plaque of rice roots subjected to differenttreatments.







Fig.3. Total As in rice roots (a), straws (b), husks (c) and grains (d) grown in
different silicon treatments (mg/kg; mean ±SD).

